# MICHIGAN DEPARTMENT OF NATURAL RESOURCES FISHERIES DIVISION 

Fisheries Research Report No. 1954
December 28, 1988

# IDENTIFICATION, MOVEMENT, GROWTH, MORTALITY, AND EXPLOITATION OF WALLEYE STOCKS IN LAKE ST. CLAIR AND THE WESTERN BASIN OF LAKE ERIE ${ }^{1}$ 

Robert C. Haas
Mary C. Fabrizio
and
Thomas N. Todd ${ }^{2}$

[^0]
#### Abstract

The harvest of walleye by sport and commercial fisheries in lakes St. Clair and Erie is under a cooperative management program involving several states and two countries. In this report we present the results of a long-term tag-recapture study as well as corroborative evidence of stock discreteness from studies of population characteristics such as growth and allelic frequencies of walleye in these waters. Walleye were tagged in the spring from 1975-87 in lakes St. Clair and Erie. Tag-recapture data indicate a general tendency for walleye to move northward after tagging. Walleye tagged in Lake St. Clair had higher recovery rates and lower survival rates than walleye tagged in Lake Erie. A reward-tag study in Lake St. Clair provided an estimate of a non-reporting rate of approximately $33 \%$ which is comparable to rates in the literature for other species. Data from the Ontario commercial (gill-net) fishery, Michigan Department of Natural Resources trap-net surveys, and sport fisheries from western Lake Erie and Lake St. Clair were analyzed with a catch-at-age model which permitted estimation of population abundance ( 12.2 to 34.5 million fish), fishing mortality rate ( 0.19 to 0.37 ), and annual survival rate ( 0.57 to 0.68 ). It appears that exploitation rates for the sport fishery in the western basin exceeded those of the commercial fishery from 1978-82. In recent years (1983-87), exploitation rates were comparable. Average abundance and catch of walleye in the western basin were 12.2 and 3.4 million fish in 1978-82; average abundance and catch in 198387 were 34.5 and 5.2 million fish. We found good agreement between the estimate of the harvest from creel surveys and that from the catch-at-age model for Lake Erie. Walleye abundance and harvest in Lake St. Clair were $10 \%$ of the values for the western basin of Lake Erie. Two discrete stocks were delineated by analysis of allelic frequencies of samples from Lake St. Clair and Lake Erie spawning populations. These two stocks are the western basin of Lake Erie and Lake St. Clair stocks. No further subdivision of stocks was possible based on the genetic analysis of 21 loci. These genetically different stocks intermix in the northern waters of this system. Based on a consideration of the results of the genetic analysis, catch-atage analysis, and tag-recapture study we recommend independent but coordinated management of the walleye populations in Lake St. Clair and Lake Erie.


## INTRODUCTION

From 1900-50 the walleye (Stizostedion vitreum) was one of the most valuable sport and commercial species in the lower Great Lakes; however, their abundance declined, apparently due to cultural and biological stresses, and reached an all-time low during the early 1960's (Regier et al. 1969). In the late 1960's fishery managers on Lake St. Clair and Lake Erie initiated a cooperative management program for Lake Erie resources. Since 1976, walleye in western Lake Erie have been managed under an interagency quota system. The Michigan Department of Natural Resources (MDNR) began a study of the Lake St. Clair walleye population in 1975 and the western Lake Erie population in 1978. This study covers the period 1975-87. Extensive sport fisheries were active on Lake St. Clair and Lake Erie during that period but no commercial walleye harvest was permitted in Lake St. Clair or in the Michigan and Ohio waters of western Lake Erie.

Extensive walleye harvests occur each year throughout the St. Clair-Detroit River system and western Lake Erie. Harvest methods are restricted to angling except in the Ontario waters of southern Lake Huron and Lake Erie. Estimates of the walleye harvest for the entire area are not available; however, the annual harvest in western Lake Erie averaged $6,873,800$ fish from 1985-87 (Lake Erie Committee, Great Lakes Fishery Commission, unpublished data). This level of harvest was roughly equal to the highest commercial harvest from Lake Erie ( 7,000 tons) taken in 1956 just prior to the walleye population collapse (Nepszy 1977). The commercial harvest of walleye from Lake St. Clair during the period 1870-1965 averaged about 50 tons (Johnston 1977). Commercial harvest of walleye in Michigan's and Ontario's waters of Lake St. Clair has been banned since 1919 and 1970, respectively. The annual walleye angler harvest in Michigan's waters of the St. Clair River, Lake St. Clair, and Detroit River in 1983 and 1984 averaged $134,143,132,454$, and 163,828 fish, respectively (Haas et al. 1985). The average walleye harvest from Ontario waters of Lake St. Clair was 109,000 fish from 1977-87 (D. MacLennan, Ontario Ministry of Natural Resources, unpublished data). The creel survey estimates for the St. Clair and Detroit rivers probably reflect about half the daytime walleye catch since there is an equally intense sport fishery in Ontario's half of these waters. In addition, the night-time harvest of walleye by anglers from the St. Clair and Detroit rivers is thought to be substantial but has not been measured.

Efficient utilization and allocation of the walleye resource requires knowledge of population distribution and abundance. The size of the walleye populations in Lake St. Clair and the western basin of Lake Erie, and the number of Lake Erie walleye that migrate into Lake St. Clair (where they are exploited under a different regime) are critical questions for quota management.

Early tagging studies showed that walleye moved within and between Lake St. Clair and Lake Erie (Wolfert 1963; Ferguson and Derksen 1971; VanVooren 1978). The tagging evidence
for migration between these waters and the decline in commercial stocks in some areas (Schneider and Leach 1977) suggested the existence of two or more genetic stocks. This prompted management agencies to examine stock structure and migration pattern, and to determine the feasibility of management based on the genetic stock concept.

Appropriately designed tag-recapture experiments provide information on survival rates and the spatial and temporal distribution of individual fish. These distribution patterns through time are indicators of population movements and mixing. Survival and exploitation rates are estimated by noting relationships between number of individuals tagged and number of tagged individuals recaptured (see e.g., Ricker 1975). In this study, we investigated nonreporting in the walleye fishery through a reward-tag experiment.

Differences in population parameters such as growth rate, recruitment, or mortality rate may be used to delineate stocks (Ihssen et al. 1981). These parameters reflect population responses to different environments and thus are not necessarily indicative of reproductive isolation (i.e., genetic stocks). However, the population parameter approach may permit delineation of stocks for management purposes and may provide corroborative evidence of stock discreteness. In this study, we examined growth rates of Lake Erie and Lake St. Clair walleye.

We also examined biochemical genetic differentiation among walleye using electrophoretic techniques and supplemented information on stock discreteness with data from a tag-recapture study. A preliminary mixed-stock sample of walleye collected during fall, 1980, in Lake Erie (Figure 1A) was screened for genetic variation at 84 enzyme loci. Ten of the resolved loci were polymorphic and the mean heterozygosity for the sample was 0.057 (J. Seeb, L. Wishard, and P. Abersold, Pacific Fisheries Research, Seattle, Washington, unpublished data). We were encouraged by the relatively high genetic variability which might allow estimation of stock structure. We collected data from spawning walleye at eight locations in Lake St. Clair and Lake Erie and examined these data for stock differences using electrophoretic and growth information.

## Study area

The study area comprises 862 km of Great Lakes waterways bounded by Lake Huron and Lake Erie (Figure 1A). We have divided it from north to south into five geographical units: southern Lake Huron, the St. Clair River, Lake St. Clair, the Detroit River, and Lake Erie. Lake Huron, the fifth largest lake in the world, is 330 km long and has a surface area of $59,674 \mathrm{~km}^{2}$. The St. Clair River is 63 km long, has an approximate flow of $5,300 \mathrm{~m}^{3} / \mathrm{s}$, and falls 1.5 m over its course (Direcki 1984a). The upper half of the river is characterized by a narrow channel with very little littoral area. Lake St. Clair is approximately 30 km long and has a surface area of $1,100 \mathrm{~km}^{2}$. The Detroit River is 51 km long, has an approximate flow of
$5,400 \mathrm{~m}^{3} / \mathrm{s}$, and falls 0.9 m over its length (Direcki 1984b). The St. Clair River has an approximate surface area of $65 \mathrm{~km}^{2} ; 58 \%$ is deeper than 6.4 m . The Detroit River has a surface area of $86 \mathrm{~km}^{2} ; 48 \%$ is deeper than 6.4 m (D. Schloesser, U. S. Fish and Wildife Service, personal communication). The Michigan-Ontario boundary roughly divides the St. ClairDetroit River system in half. The western basin of Lake Erie has a surface area of $3,276 \mathrm{~km}^{2}$. Major walleye spawning tributaries have been identified: Ontario's Thames River on Lake St. Clair (Ferguson and Derksen 1971) and Ohio's Maumee River on Lake Erie (VanVooren 1978; Figure 1A). There are no walleye spawning grounds in Michigan's waters of Lake St. Clair or Lake Erie where successful reproduction has been verified in recent times. Lake St. Clair is characterized by a diverse plant and invertebrate animal community which, compared to Lake Erie, has shown relatively little evidence of negative impacts of enrichment (B. Manny, U. S. Fish and Wildlife Service, personal communication).

## METHODS

## Net samples

Trap nets set in the spring in Lake St. Clair and Lake Erie provided abundance data on age 2 and older walleye. Gill nets set in the fall in Lake Erie provided data on abundance of yearling and older fish. Gill nets typically provide indices of relative year-class strength (Willis 1987). Impoundment gear (trap net) is generally considered to be superior for relative abundance studies (Yeh 1977; Craig 1980); however, traps must be fished for extended time periods which is expensive. We examined the relative year-class strength indices from the two gear types for Lake Erie because gear selectivity influences the size distribution of the sample.

Trap nets were used to capture fish for tagging and to provide an index of relative abundance, catch per unit of effort (CPUE; number caught per 24 hours). This assumes that CPUE is linearly related to fish abundance and that a percent change in CPUE will reflect the same percent change in abundance (Bannerot and Austin 1983). We captured walleye during spring in Lake St. Clair and Lake Erie with 6 -foot-deep trap nets fished at the same locations each year (Figure 1B). Five nets were fished throughout each sample period and were normally tended on all weekdays, weather permitting. We tried to obtain a minimum of 50 net lifts for each period. The Lake St. Clair surveys were carried out from 1975-85 and the Lake Erie surveys from 1978-87. The average Lake Erie sampling period began on April 17 and lasted 20 days; the average Lake St. Clair sampling began on May 23 and lasted 29 days.

The entire catch from each trap net was identified and enumerated. Growth data and age samples (scales) were collected from walleye and other species. The maximum time between net lifts was 72 hours; the majority was lifted after 24 hours. We compared the fish communities of Lake St. Clair and western Lake Erie with the Shannon-Weiner diversity index
(Vandermeer 1981) calculated from the trap-net data. This index measures the evenness of the distribution of individuals within species. The Shannon-Weiner function is given by:

$$
\mathrm{H}=\sum \mathrm{P}_{\mathrm{i}} \ln \left(\mathrm{P}_{\mathrm{i}}\right)
$$

where $P_{i}$ is the fraction of all individuals in the community contained in species $i$ (Vandermeer 1981). The value of H is largest when the number of species is greatest and when individuals are apportioned most evenly among species.

We fished multi-filament, graded-mesh gill nets at two stations in Lake Erie (Figure 1B) in October from 1978-87 as part of the interagency yearling-walleye index program. Replicate sets were made each year with gangs of nets, 1.83 m deep, each consisting of seven 30.48 m long panels that ranged from 51 to 127 mm stretched mesh measure by 13 mm intervals. All walleye captured in gill nets each year were sampled for age and growth data.

## Tag-recapture study

Walleye were tagged by MDNR personnel during the spring trap-net surveys in Lake St. Clair and western Lake Erie. We did not consider these tagged populations as spawning stocks because fish were tagged after spawning had occurred; none of the known walleye spawning grounds were located in Michigan's waters. The Lake St. Clair tagging station was located approximately 47 km northwest of Ontario's Thames River spawning ground and 129 km north of the Maumee River spawning ground (Figure 1B). The Lake Erie tagging station was located near Monroe, Michigan, 24 km north of Ohio's Maumee River spawning site.

Fish were removed from the trap nets and immediately placed in an on-board live tank equipped with continuously circulating lake water. Fish were removed individually from the live tank and tagged without anesthesia before release at the net location. Total length measurements were made on all tagged fish, while total weight measurements and scale samples were taken from portions each year varying from $36 \%$ to $100 \%$ of the total number tagged. When scale samples were taken, all fish from that trap net were processed. Net lifts were sampled throughout the survey period for age data. All fish were tagged with size 10 or size 12 monel metal strap tags affixed by overlapping the tag snugly around the dentary bone. The tags were inscribed with the local MDNR address and individual tag number. Tag-recapture data were solicited from anglers and commercial fishermen on a voluntary basis.

We tagged 11,876 walleye at the Anchor Bay site, Lake St. Clair, from 1975-85 and 17,957 walleye at the Monroe site, Lake Erie from 1978-87.

The time of tagging was an important variable in defining the movement of the two stocks and the rate of mixing. The Lake Erie tagging was carried out in April each year; most fish were captured immediately after spawning. The Lake St. Clair tagging followed; most fish
were tagged during the last week of May and first week of June. The average Lake Erie walleye was tagged about 36 days earlier than the average Lake St. Clair fish. Walleye tagged in Lake Erie were most likely all Lake Erie stock primarily from the Maumee River because of the proximity of tagging date to spawning times; other stocks would not have had time to migrate that far. However, the walleye tagged in Lake St. Clair might have been either Lake St. Clair or Lake Erie stock since ample time was available for migration and mixing.

Tag-recovery data were summarized by location and Julian date. Dates of tagging and tag recovery for recaptured walleye were coded by the Julian calendar and thus were independent of the calendar year. This permitted calculation of the extent of dispersal from the tag sites and time-at-large. Numbers of tags from each tag site, recovered within a specific area, were compared to provide estimates of stock mixing. Nonparametric statistics were used to test for differences between median recapture dates.

Recently, a generalized stochastic model (and sophisticated computer algorithm) became available for the analysis of data from tagging experiments (Brownie et al. 1985). This model (hereafter referred to as the BROWNIE model) provides maximum likelihood estimates of recovery and survival rates which are unbiased. Total mortality rate (natural logarithm of survival rate) may be partitioned into fishing and natural mortality rates if an estimate of the tag reporting rate is available (Horsted 1963). This is because tag recovery rate is a product of the exploitation rate and the reporting rate (Krementz et al. 1987). (Brownie et al. [1985] refer to the recovery rate as the reported exploitation rate.)

The BROWNIE model was used to analyze the results of the tag-recapture experiments in Lake St. Clair and Lake Erie. We estimated mean annual survival, recovery, and adult life span and these values were compared statistically using the recommended $z$ test (Brownie et al. 1985).

In many studies the reporting rate is assumed to be $100 \%$, that is, all of the tags recovered by the fisheries are seen and subsequently reported. If $100 \%$ reporting is assumed, then the recovery rate is an estimate of the exploitation rate. More likely, reporting rate is less than $100 \%$ and may vary over time (Rawstron 1971), space (Chadwick 1968; Henny and Burnham 1976; Reeves 1979; Green et al. 1983), or other factors (Rawstron 1971; Green et al. 1983).

If an estimate of the exploitation rate is available, then fishing mortality rate may be estimated. However, fishing mortality rate is underestimated whenever the assumption of complete reporting is violated. As reporting rate decreases, the relative error of the recovery rate (and hence, the relative error of the exploitation rate) increases (Conroy and Williams 1981; Figure 2). For example, if $80 \%$ of the tags are recognized and reported, the relative error of the recovery rate is $25 \%$. Clearly, estimation of the exploitation and fishing mortality rates will be most reliable when reporting rates are high.

High reporting rates are difficult to ensure. Rewards, ranging from money to books to chances in a lottery, have been offered for the return of tags. Presumably the monetary reward or prize is a further incentive to the angler or commercial fisherman to report the catch. Holt (1963) suggested "planting" of marked fish in the catch of commercial fishing vessels as a way to estimate non-reporting. Margetts (1963) used this method to study non-reporting in the British cod fishery in the Barents Sea. Green et al. (1983) inserted (used) internal anchor tags in the catch of marine recreational anglers in Texas to examine non-reporting for seven species.

A reward-tag study was carried out on Lake St. Clair in 1981, 1982, and 1983 to provide an estimate of the non-reporting rate for traditional non-reward tags. Funds to pay rewards were solicited from local conservation organizations. Reward tags of four denominations, $\$ 2.00, \$ 4.00, \$ 6.00$, and $\$ 8.00$, carried a reward inscription, such as "Reward $\$ 8.00$ ", and returns from these tags were used to examine the effect of monetary value on angler cooperation. Every third walleye received a reward tag. The different reward denominations were applied in repeating sequence from $\$ 2.00$ through $\$ 8.00$ so that no denomination bias was introduced.

## Relative abundance

Relative year-class strength for Lake Erie walleye was calculated by estimation of the mean CPUE for ages 1 to 5 for 1977-82 year classes. (All five age groups were not represented in years prior to 1977 and after 1982.) Catch of age 6 and older fish was low and did not significantly contribute to the overall catch for a particular year. Gill-net and trap-net indices were compared with a nonparametric test of association (Hotelling-Pabst test for significance of Spearman's rho). This test does not require assumption of bivariate normality (Conover 1980). The null hypothesis tested was that gill-net and trap-net indices for 1977-82 year classes are mutually independent, i.e., the two gear types do not provide similar estimates of relative abundance.

We used the catch-at-age model, CAGEAN (Deriso et al. 1985), to estimate current and historical walleye stock sizes in Lake St. Clair and the western basin of Lake Erie. This model uses fishing mortality and catch-at-age data to arrive at stable and reliable estimates of current stock size. It is an improvement of the traditional virtual population analysis since multiple gear types and auxiliary information on fishing effort are explicitly considered in the model. The Marquadt algorithm is used to solve a system of nonlinear equations and to provide leastsquare estimates of model parameters. We used the IBM personal computer (PC) version of CAGEAN; this "reads in" gear-specific catch-at-age, fishing effort, and weight data and estimates age-specific abundance, catch, fishing mortality, selectivity, and catchability (CAGEAN-PC User Manual 1987).

We felt this age-structured program would be useful because it calculates stock abundance estimates from historical catch and effort data (including auxiliary survey data) as well as estimates of mortality and exploitation rates. We compared CAGEAN-derived estimates of mortality with estimates from tag analyses. Angler catch and effort data from creel surveys were combined with the trap-net survey data to produce comparable catch and stock abundance estimates for Lake St. Clair and Lake Erie walleye.

## Growth

The western basin population has undergone significant change during the past 15 years generally from low to high abundance. There is evidence that growth has responded to increased density by decreasing in recent years. The western basin population is also under the most critica! management program because it is shared by at least two states and one Canadian province and is sought by expanding sport fisheries and efficient commercial gill-net fisheries.

Some of the walleye collected by trap net were sampled for scales with the intention of age determination and subsequent growth analyses. Mean length-at-age was estimated for Lake St. Clair and Lake Erie walleye sampled from 1975-87. Because growth of Lake Erie walleye was suspected to differ before and after 1982, mean lengths-at-age were computed for the periods 1978-82 and 1983-87. Growth was described by the von Bertalanffy model,

$$
L_{t}=L_{\max }\left[1-\exp \left\{-\mathrm{k}\left(\mathrm{t}-\mathrm{t}_{\mathrm{o}}\right)\right\}\right]
$$

fit to mean length data for age 1-9. Model parameters were estimated with a nonlinear least squares method (FISHPARM, Prager et al. 1987), the preferred method of fitting this model (Vaughan and Kanciruk 1982).

We examined the relationship between length and age with regression analysis for Lake Erie and Lake St. Clair walleye. Because we found different relationships for the two groups (see Results section) we did not use the regression model for size-at-age comparisons. Instead, comparisons of mean length-at-age were made with a $t$-test.

Growth differences of year classes were examined by computing relative growth indices:

$$
\operatorname{RGI}_{j}=\sum_{i=1}^{\mathrm{k}}\left[{\overline{\mathrm{~L}(\mathrm{t}})_{j}}^{\mathrm{j}} \overline{\mathrm{~L}(\mathrm{t})}\right]
$$

where

$$
\begin{aligned}
\mathrm{RGI}_{\mathrm{j}} & =\text { the relative growth index of jth year class } \\
\mathrm{k} & =\text { number of age groups } \\
\overline{\mathrm{L}(\mathrm{t})} & =\text { mean length-at-age over all year classes } \\
\overline{\mathrm{L}(\mathrm{t})} & \mathrm{j}
\end{aligned}=\text { mean length-at-age of jth year class }
$$

The purpose was to relate relative growth of a year class with its year-class strength (as measured by the rank transform of trap-net CPUE).

## Genetic analysis

Walleye were collected during spawning at six sites in Lake Erie and two in Lake St. Clair (Figure lA) with gill nets or by electrofishing. Mixed-stock samples of approximately 200 fish each were collected with gill nets from Anchor Bay, Lake St. Clair, in November 1983 and 1984. Whole fish were placed on ice immediately and taken to the laboratory where they were kept frozen at $-20^{\circ} \mathrm{C}$. Fish were thawed in the lab to extract muscle and liver tissues. Tissues were then ref rozen at $-80^{\circ} \mathrm{C}$ for later electrophoretic analyses. Total length, total weight, gonad weight, age, sex, and maturity were collected from all walleye sampled for genetic information. Somatic weight was calculated by subtraction of gonad weight from total weight.

Horizontal starch-gel electrophoresis was employed to separate the various enzymes. The staining procedures were modified from Allendorf et al. (1977) and Harris and Hopkinson (1977). Stains with agar overlays were made according to Todd (1983); we used the nomenclature for allelic variants suggested by Allendorf and Utter (1979).

BIOSYS-1, a FORTRAN program by Swofford and Selander (1981) was used to analyze the allelic frequency data. Measures of genetic variability include allelic frequencies and mean heterozygosity. Polymorphic loci were tested for conformance to Hardy-Weinberg expectations of random mating and inter-populational heterogeneity of allelic frequencies by means of Chisquare tests. Stock composition of the fall samples from Lake St. Clair was obtained by an iterative procedure which calculated suspected genotypic frequencies from the observed allelic frequencies of the Lake St. Clair and Lake Erie samples.

## RESULTS

## Net samples

Trap-net catches indicated the fish communities in Lake St. Clair and western Lake Erie were similar with respect to kind and number of species (see Appendices 1,2 , and 3 for a species list and annual trap-net CPUE). The Shannon-Weiner index value calculated from the Lake St. Clair trap-net data was 2.10 and from the Lake Erie trap-net samples, 1.24. Since the number of species was nearly identical (Lake St. Clair, 39 species; Lake Erie, 40 species), the rather large difference observed was most likely due to the relatively uneven distribution of individuals within species in the Lake Erie samples. Mean trap-net CPUE values for 20 of the most prevalent species in Lake St. Clair and Lake Erie are shown in Figure 3. Trap nets in Lake Erie sampled 3.34 times as many fish; however, those individuals were mostly yellow perch, which had a CPUE 50 times greater than in Lake St. Clair. We feel that the Lake Erie fish community is unbalanced relative to that in Lake St. Clair; this may have implications for walleye growth and year-class success in Lake Erie (see Discussion section).

Trap nets captured considerably more walleye in Lake Erie (mean CPUE was 22.9) than in Lake St. Clair (mean CPUE was 8.5). The Lake St. Clair and Lake Erie trap-net CPUE of walleye age groups for each sample with corresponding effort data are given in Appendices 4 and 5. It is important to note that Lake Erie walleye were almost fully recruited to the trap nets at age 2 , while the Lake St. Clair walleye were not fully recruited until age 3. The fall gillnet CPUE for walleye by age group for each season in Lake Erie is given in Appendix 6.

The Hotelling-Pabst test of the null hypothesis of no association between gill-net and trap-net abundance indices (Table 1) indicated that the association was significant ( $r_{s}=0.899$, $N=6, P<0.01$ ). Thus, the relative year-class composition of the walleye population was well represented in the trap-net samples taken for the tag-recapture study. A similar analysis of ages 2-5 from 1976-82 indicated that gill-net and trap-net samples yielded similar relative abundance estimates ( $r_{s}=0.964, N=7, P=0.001$ ). Inclusion of age 1 in the mean catch per unit effort decreased the strength of the association, but not significantly. Graded-mesh gill nets, designed to capture small fish, provided more efficient samples of age-l fish than did trap nets.

## Growth

The von Bertalanffy model was a reasonable description of growth for Lake St. Clair and Lake Erie walleye (Figure 4). Growth constants were similar ( $k=0.104$, Lake St. Clair; $\mathrm{k}=0.102$, Lake Erie) although the asymptotic length values were not; Lake Erie walleye growth is characterized by a higher asymptotic length ( $\mathrm{L}_{\max }=902 \mathrm{~mm}$ ) than growth of Lake St. Clair
walleye ( $\mathrm{L}_{\max }=854 \mathrm{~mm}$ ) (Figure 4). Von Bertalanffy model parameters fit to two data sets from Lake Erie (1978-82 and 1983-87) indicate greater asymptotic lengths in the second period ( 689 mm versus 781 mm ). There is some indication, then, that growth of Lake Erie walleye changed from the period 1978-82 to 1983-87.

Linear models were employed to describe length-at-age for Lake Erie $(N=12,312)$ and Lake St. Clair ( $N=7,237$ ) walleye taken by trap net. The two-parameter models investigated were:
(I) Length $=\mathrm{a} \cdot($ age $)+\mathrm{b}$
(II) $\log ($ Length $)=a \cdot($ age $)+b$

Although both regressions were significant, Model I exhibited the best fit as determined by $r$-square values (Model I $r^{2}=0.78$ and 0.80 for Lake Erie and Lake St. Clair; Model II $r^{2}=0.75$ and 0.77 ). This linear model accounted for the greatest amount of variability in observed length. The slopes of the Model I regressions for Lake Erie and Lake St. Clair walleye were significantly different ( $F=888.0, \mathrm{df}=1,19,548 ; P<0.01 ; F$-statistic for difference in slopes, Sokal and Rohlf 1981), implying that for a given age, total lengths are significantly different. Mean lengths for age 2 to age 10 were compared with a $t$-test (Table 2). At ages 3, 4, 5, 6, and 8 walleye from Lake Erie were significantly larger than the same age fish from Lake St. Clair. Lake St. Clair walleye were significantly larger than those from Lake Erie only at age 2 , although this difference was small and biologically insignificant ( 340 versus 337 mm ).

For Lake Erie walleye (1972-84) the 1972 year class exhibited the highest overall relative growth (189.4); the 1977 and 1978 year classes had the lowest ( -115.5 and -120.7 , respectively) (Table 3). The trap-net index of relative abundance for the fast-growing 1972 year class was low; the slow-growing 1977 year class had the second highest relative abundance index. However, this apparent inverse relationship between year-class strength and relative growth was not consistent for all year classes (Figure 5). The linear relationship accounted for only $36 \%$ of the variation in the relative growth index ( $r=0.597, P<0.05$ ). For Lake St. Clair walleye (1964-82), the lowest relative growth index was -61.8 (1978 year class, rank $=12$ th highest of 19 year classes) (Table 4; Figure 6). The 1969 year class exhibited the greatest relative growth index (255.0) and was 14th in rank abundance. The linear relationship of relative growth to rank abundance was not significant, i.e., the correlation coefficient was not significantly different from zero ( $r=0.241, P>0.05$ ). The variation in observed relative growth could not be adequately modeled with rank abundance.

The abundance of a particular year class may influence growth of subsequent year classes. Therefore, we examined the cumulative effects of year classes on growth by computing 3 -year and 2 -year running averages of trap-net CPUE. There was no significant linear relationship between 3 -year or 2 -year running averages of CPUE with relative growth in Lake

Erie walleye (3-year average: $r=0.238, P>0.05$; 2-year average: $r=0.339, P>0.05$ ). This attempt to relate relative growth to 3- and 2-year running averages accounted for $34 \%$ (3-year average) and $24 \%$ (2-year average) of the variation in relative growth of Lake St. Clair walleye. Variability in growth of walleye year classes was best modeled by cumulative effects of year-class abundances in Lake St. Clair populations whereas, the variability in growth of walleye year classes in Lake Erie is better described by the relative abundance of individual year classes.

## Tag-recapture study

A total of 1,333 Lake St. Clair tags and 1,159 Lake Erie tags were eventually recovered by fishermen and voluntarily reported. The lower reporting rate for the Lake Erie tags probably reflects a lower exploitation rate compared to Lake St. Clair, instead of poorer cooperation from Lake Erie fishermen.

The major portion of the tag recoveries were reported by anglers; $10 \%$ of tag returns were from commercial fishermen. There appears to be ample angling harvest throughout the area to provide enough voluntary tag recoveries to adequately monitor movements of the tagged stocks.

The areal distribution of all recaptures of Lake St. Clair and Lake Erie tags are summarized by recovery area in Figure 7. This figure shows the distribution of recoveries expressed as percent of total recovered within each lake. Lake Erie walleye show a consistent and relatively strong tendency to move northward into the Detroit River, Lake St. Clair, and as far north as Saginaw Bay, Lake Huron ( 170 km ). Lake St. Clair walleye also show a strong tendency to migrate northward (Figure 7). Only $3 \%$ of the Lake St. Clair tags were recovered in Lake Erie indicating a very low rate of movement southward. There is also some evidence to suggest that Lake Erie recoveries of walleye tagged in Lake St. Clair tend to be walleye of Lake Erie origin that had originally migraced to Lake St. Clair early enough to be intercepted by the Lake St. Clair tagging effort. The evidence was obtained from comparison of the mean Julian date of recapture in Lake Erie of fish tagged in Lake St. Clair with the mean date of recapture in Lake Erie for all Lake Erie tagged walleye.

Twenty-nine percent of all recoveries of Lake Erie tags were reported from the Detroit River and north (Table 5). This ratio did not vary between years in spite of dramatic increases in Lake Erie angling harvest and walleye abundance.

The monthly distribution of tag returns from Lake St. Clair and Lake Erie are shown in Figures 8 and 9. On-site creel surveys of the angler catch indicated that tag recovery patterns were similar to patterns of harvest. Over $78 \%$ of the walleye tagged in Lake St. Clair were recovered from Lake St. Clair in May through August. Data from a creel survey of Michigan waters of Lake St. Clair in 1983 and 1984 indicated that $88 \%$ of the boat harvest of walleye occurred during that same period (Haas et al. 1985). Eighty-one percent of the Lake Erie tags
were reported from Lake Erie waters in May to July. Creel surveys of Michigan waters of Lake Erie showed that $96 \%$ of annual walleye harvest in 1986 and 1987 was taken in May to July (G. Rakoczy, Michigan Department of Natural Resources, unpublished data).

Recapture dates within specific lake areas provided some of the most useful information for stock management purposes. We had postulated that tagged walleye returned to their natal spawning grounds (a behavior referred to as homing) and that individual fish repeated their migration route each year in a circular pattern. Peak walleye spawning activity has been observed around April 3 in the Thames River (D. MacLennan, Ontario Ministry of Natural Resources, personal communication) and around April 1 in the Maumee River (C. Baker, Ohio Department of Natural Resources, personal communication). Most of our Lake Erie walleye were tagged within 3 weeks of peak spawning so they probably were part of the Maumee River stock. Since tagging was protracted in Lake St. Clair, those fish might have originated from either the Lake St. Clair or the Lake Erie stock.

We analyzed tag data according to the postulate that any particular tagged walleye would return to the same location on a particular Julian day; for example, back at the original tag site on subsequent tagging anniversaries. With this postulate in mind, we grouped the date information by lake of tagging and lake of recapture and made comparisons of median Julian days using the Mann-Whitney $U$ and Wilcoxon two-sample tests. Nonparametric statistics were used because we found that variances were not homogeneous. Median Julian dates for all possible combinations of tagging and recovery are shown in Table 6. Walleye tagged in Lake St. Clair were recovered on day 155 (median recovery day) in Lake Erie; this was significantly earlier ( $P=0.003$ ) than recovery day 180 for walleye tagged and recovered in Lake Erie. This indicated that some of the walleye tagged in Lake St. Clair and subsequently recaptured in Lake Erie may have originated from the Lake Erie stock. We postulate that their period of vulnerability to capture in Lake Erie was much shorter and earlier in the year as they were migrating to and from their natal Lake Erie spawning grounds. Apparently, very few tagged walleye from Lake St. Clair migrated to Lake Erie since the median recovery date in Lake Erie should have been the same as, or later than, the median recovery date of the Lake Erie stock. In other words, if numerous walleye tagged in Lake St. Clair had subsequently migrated to Lake Erie, then their median angler recovery day in Lake Erie should have been June 29 (day 180) or later, consistent with their probable vulnerability to fishing in Lake Erie.

Recovery data from the north half of the St. Clair River were evaluated for estimates of the rate of mixing of the Lake St. Clair and Lake Erie stocks. We chose this recovery site because the narrow river is characterized by substantial and consistent angling pressure, and angler exploitation of walleye and other species is higher in this area than elsewhere in the system. An estimate of the average annual daytime harvest of walleye in Michigan waters of the upper St. Clair River from 1983-84 was 120,000 fish (Haas et al. 1985). Seventeen percent
of all recoveries of Lake St. Clair tags and $6 \%$ of all recoveries of Lake Erie tags were reported from this area.

We compared the distribution of recovery dates in the upper St. Clair River for the Lake St. Clair and Lake Erie stocks to see if they were similar. The monthly distribution, as percent tags recovered in the upper St. Clair River, is shown in Figure 10. Both stocks showed the highest recovery rates in June and July with Lake Erie walleye slightly skewed toward earlier recaptures. The distributions of Julian recovery dates for the two groups of tagged fish were identical (Mann-Whitney $U$-test; $P=0.33$ ).

The overall ratio of Lake Erie tags to Lake St. Clair tags recovered in the upper St. Clair River was $68: 226$. We feel this ratio represents the best estimate of stock mixing: $23.1 \%$ of the upper St. Clair River walleye population was Lake Erie stock. This estimate agrees favorably with the estimate derived from genetic analysis of walleye sampled from Lake St. Clair (see section on Mixed-fishery analysis, below). This consistency invites speculation on the use of tag-recovery data for estimation of stock contributions. While it is true that there are many assumptions (e.g., equal distribution of fishing pressure throughout the study area for a particular period of time, equal non-reporting bias, equal catchability, complete mixing of the tagged group of fish with non-tagged individuals, etc.), in general, tag recovery ratios may provide a preliminary assessment of the extent of stock mixing. We calculated these ratios for Lake Erie, Lake St. Clair, and the Detroit River from total tag returns for the period 1975-87 and thus these estimates represent annual average stock contributions. The Lake Erie stock comprised approximately $23 \%, 8 \%, 95 \%$, and $70 \%$ of the walleye populations in upper St. Clair River, Lake St. Clair, Lake Erie, and the Detroit River (Appendix 7).

Walleye survival rate estimates were generated from non-reward tag-recovery data using the IBM mainframe computer version of the BROWNIE model. The mean survival and tag recovery rate estimates for Lake St. Clair and Lake Erie are shown in Table 7. These estimates were compared using the $z$ test which indicated that the survival rate was significantly greater for Lake Erie walleye than for walleye from Lake St. Clair ( $z=2.911, P<0.002$ ). The mean life span for adult walleyes in Lake Erie was significantly longer ( $z=2.817, P<0.002$ ). The mean recovery rate for Lake Erie tagged walleye was also significantly less ( $z=7.803$, $P<0.001$ ) than the estimated recovery rate for the Lake St. Clair walleye; however, this comparison is valid if angler reporting behavior, which is unknown, is the same for both lakes.

## Reward-tag study

A reward-tag study on walleye was carried out on Lake St. Clair from 1981-83 to provide an estimate of the frequency of non-reporting, i.e., tag recovery not followed by tag reporting. There were 909 walleye tagged with reward tags and 2,043 tagged with non-reward tags during the 3 -year reward study on Lake St. Clair. The number of walleye tagged with non-reward
tags and the number tagged with each reward denomination from 1981-83 are listed in Appendix 8. The numbers reported to MDNR by fishermen from time of tagging through 1987 are also given.

There was an overall ratio of 1.6 walleye reward tags reported for every non-reward tag. These data were examined further to determine whether the recovery frequency for reward tags was significantly greater than the non-reward tags. The recovery data derived from the three tagging years were combined for each of five sequential recovery seasons, without regard to calendar year, since the ratio of one reward to two non-reward tags was maintained throughout the 3 -year tagging period. This resulted in five sequential recovery periods for each tagged cohort; 5 years was essentially the maximum survival period for any tagged fish. The percent of walleye with reward and non-reward tags recovered each season is shown in Figure 11. The observed recovery frequency of non-reward tags was used to calculate the expected values for the reward tags and frequencies of reward recoveries were higher than non-reward frequencies during all recovery seasons. The reporting frequency of walleye reward tags was different from the frequency of non-reward tags ( $\chi^{2}=47.2, P<0.005$ ).

The reporting rate for walleye reward tags may be related to the monetary value of tags. The frequencies of recovery of $\$ 2.00$ tags was compared to the frequencies for each of the larger denominations. The recoveries of the $\$ 4.00$ ( $20.0 \%$ ), $\$ 6.00$ ( $19.1 \%$ ), and $\$ 8.00$ ( $21.0 \%$ ) tags were slightly higher than the $\$ 2.00$ ( $15.6 \%$ ) tags but did not vary among themselves. None of the larger denominations showed a significantly higher recovery rate compared to the $\$ 2.00$ tags ( $\$ 4.00 \chi^{2}=4.1, P>0.10 ; \$ 6.00 \chi^{2}=2.8, P>0.10 ; \$ 8.00 \chi^{2}=4.5, P>0.10$ ).

Although reward tags of four denominations ( $\$ 2.00, \$ 4.00, \$ 6.00$, and $\$ 8.00$ ) were used in this study, there was no clear indication that an increase in reward was associated with an increase in reporting rate. Therefore, all recapture data were pooled across reward denominations. Rawstron (1972) found no difference in reporting of $\$ 1.00$ and $\$ 5.00$ reward tags from largemouth bass. Furthermore, he interviewed anglers and found that a $\$ 5.00$ reward was insufficient inducement (Rawstron 1971). Our results indicate that there was no appreciable difference in reporting of tags, irrespective of the value. Recovery rates were significantly different for reward (0.189) and standard or non-reward (0.120) tags pooled across years (two-tailed $t$-test for significant difference in ratios, Cochran 1977; $t=21.805$, $\mathrm{df}=2, P<0.05$ ). Reward tags were recovered at a significantly higher rate (approximately $7 \%$ higher) than the standard tags.

Unbiased estimates of survival rate from the tag-recapture model (Brownie et al. 1985) may be used to estimate fishing mortality rate if the reporting rate is known. A single estimate of the reporting rate, the ratio of reward-tag to standard-tag return rate, is needed. This ratio may be highly variable for a given year after tagging as well as for a given cohort of tags (Table 8). We examined these ratios over time and between cohorts of tags (1981 versus 1982,
e.g.). The difference in reporting rate in the first and second years (first year is year of tagging ) is significant for reward and standard tags ( $t=19.021$, year $1, \mathrm{df}=2 ; t=5.463$, year 2 , $\mathrm{df}=2 ; P<0.05$ ). These differences were not significant in the third, fourth, or fifth years. This implies that reporting rates are essentially the same in later years regardless of the type of tag (reward or standard). However, the rate of return of standard tags is significantly less than the rate of return of reward tags for the first and second years; therefore, tag recovery data are biased by non-reporting in the first 2 years, but generally not thereafter. Variation in recovery rates, due to reporting rate or other factors, does not bias the survival rate estimates from the tag-recapture model when survival rates are constant for all individuals (BROWNIE Model 1) (Pollock and Raveling 1982).

The $95 \%$ confidence interval for the ratio of reward-tag to standard-tag return rate in the year of tagging is 0.88 to 2.34 . In the following year it is 0.91 to 1.59 ( $95 \%$ confidence interval for ratio of two ratios, Cochran 1977). These two estimates are not significantly different and a combined estimate for the ratio is 1.48 (ratio of reward-tag return rate to standard-tag return rate). The reporting rate is estimated by the inverse of this ratio, i.e., it is the ratio of the standard-tag to reward-tag return rate. Assuming all the reward tags are reported, the reporting rate for the walleye fishery in Lake Erie-Lake St. Clair is 0.674 (nonreporting rate is $1-0.674=0.326$ ). If the assumption of $100 \%$ reporting does not hold, then this estimate is biased. For example, if $80 \%$ of the reward tags were reported, the bias would be $25 \%$ or 0.169 (reporting rate may range from 0.505 to 0.843 ). Clearly, the accuracy of the reporting rate is contingent upon the $100 \%$ reporting assumption. However, numerous tagging studies indicate that this assumption may not be valid (e.g., Mullan 1959, Paulik 1963, Rawstron 1971, Green et al. 1983). Non-reporting rates vary from $15 \%$ to $77 \%$ in fish tagging studies and from $50 \%$ to $60 \%$ in waterfowl and bird banding studies (Table 9). Our estimate of $33 \%$ is within the range of published non-reporting rates.

## Genetic analysis

Of 21 loci assayed, only four exhibited polymorphisms such that the frequency of the most common allele did not exceed 0.95 in any one population (Tables 10 and 11). These loci were ADH-1, GMP-3, IDH-1, and MDH-3. Mean heterozygosity ranged from 0.076 with standard error (SE) of 0.035 in the Kelley's Island sample to 0.081 ( $\mathrm{SE}=0.037$ ) in the Sandusky River sample with most of the others at 0.080 ( $\mathrm{SE}=0.037$ ). None of the samples differed significantly for this measurement. The several stocks examined in this study were similar in the amount of genetic variability measured and in shared polymorphisms.

## Within-lake heterogeneity

Walleye stocks in Lake Erie did not appear to have signif icant genetic heterogeneity and thus, individual stocks could not be delineated (Table 10). However, analyses of the pooled data from the six Lake Erie localities revealed significant departures from random mating expectations at IDH-1 and MDH-3 (Table 11). These were due to excess heterozygotes which suggest that more than one spawning stock was represented in the pooled sample.

Spawning walleye from the Clinton and Thames rivers in Lake St. Clair appeared to be genetically identical to one another (Table 12). In addition, there was no significant heterogeneity between year classes from these two rivers which was corroborated by pair-wise contingency Chi-square analyses at all loci (Table 12). Based on genetic data, there seems to be no evidence to support the hypothesis of separate stocks in the Clinton and Thames rivers.

## Between-lake variability

We found significant stock differentiation between the Lake Erie and Lake St. Clair walleye based on three loci (Table 11). Only ADH-1 exhibited no significant heterogeneity among the four most polymorphic loci. The dendrogram based on Nei's (1978) minimum genetic distance was significant at the level separating the Lake St. Clair populations from those of Lake Erie (Figure 12). We concluded that the Lake St. Clair and Lake Erie walleye stocks represent distinct gene pools and that a measure of the extent of mixing would be feasible.

## Mixed-fishery analysis

The stability of the allelic frequencies and the magnitude of the genetic differences between Lake Erie and Lake St. Clair walleye stocks permitted measurement of the proportional contribution to mixed-stock samples taken in Lake St. Clair in the fall. In 1983, the mixture was estimated to contain $92.1 \%$ Lake St. Clair fish and $7.9 \%$ Lake Erie fish ( $N=217$; standard deviation (SD) $=19.5$ ). In 1984, the mixture was estimated to contain $68 \%$ Lake St. Clair fish and $32 \%$ Lake Erie fish ( $N=253$; SD=19.8). Simulations of mixed salmon fishery data have shown that the actual proportion was contained within one standard deviation for sample sizes of 200 or more fish (Milner et al. 1983). Therefore, a reasonable expectation is that Lake Erie fish made up $0.0 \%$ to $27.4 \%$ of the November 1983 sample and $12.2 \%$ to $51.8 \%$ of the November 1984 sample. The overlap of the confidence intervals for the two estimates suggests that stock composition in the 2 years was similar; therefore, a combined estimate was calculated: 76.9\% Lake St. Clair fish and 23.1\% Lake Erie ( $N=470$; $\mathrm{SD}=14.1$ ). These were the
same proportions estimated from analyses of tag-recapture data from the northern half of the St. Clair River.

Total length and somatic weight data from the Thames River and Maumee River samples used in the genetic analysis were compared for differences in population growth to further investigate stock structure. The data were analyzed by single age/sex categories to eliminate the potential for bias. Three cohorts contributed individuals to each age/sex group. Growth differences between year classes were probably not a significant factor since the same 3 years of sampling contributed data to both sites. The two-sample $t$-test was used to compare population means and a significant $t$-statistic ( $P<0.05$ ) indicated that the two means were not identical. Mean length and somatic weight for the Maumee River population were significantly greater than the Thames River population in most of the comparisons (Tables 13 and 14). All of the comparisons with large sample sizes indicated that the Maumee River stock had significantly $(P<0.01)$ greater growth, in terms of both length and somatic weight. Regression analyses were used to examine differences in the length-weight relationships for males and females from the two sampling sites. Linear regressions of $\log _{e}$ length (mm) versus $\log _{e}$ somatic weight (g) for the two spawning samples were significantly different for males ( $F=26.8, P<0.01$ ) and females ( $F=22.3, P<0.01$ ). Maumee River walleye had significantly heavier bodies at a given length compared to the Thames River fish. Analyses of the growth data suggest that the Lake St. Clair and Lake Erie walleye belong to distinct stocks, and this conclusion corroborated the genetic evidence.

## Abundance estimates: CAGEAN model

Abundance of walleye in the western basin of Lake Erie was estimated with the CAGEAN catch-at-age model (Deriso et al. 1985). We used the CAGEAN three-gear catch-at-age model which allows survey data to be combined with data from multiple gear fisheries. Deriso et al. (1985) found that bias was substantially reduced when auxiliary information, such as effort data and survey catches, was included in the analyses. We combined western basin walleye catch and effort data from the commercial and sport fisheries for 1978-87 (unpublished Ontario and Ohio fishery data obtained through the Walleye Task Group, Lake Erie Committee, Great Lakes Fishery Commission) with MDNR spring trap-net data. Walleye growth decreased during the 10 -year period, such that recruitment to the survey trap nets was delayed by 1 year for some year classes. The data were somewhat arbitrarily grouped into two time periods, 1978-82 and 1983-87, because we speculated that the decrease in growth may have affected catchability coefficients and age-specific selectivities. Similar changes in recruitment probably took place in the sport and commercial fisheries. We set the full recruitment age to three for the sport and commercial fisheries during the early period, and to four during the later period. The full recruitment age was 1 year less for the trap-net samples. We set natural
mortality to 0.2 ; there were no data to produce a reliable estimate. The analyses of tag recovery data suggest, however, that natural mortality has been higher than 0.2 since tag reporting rate has been quite low.

The CAGEAN estimates of mean instantaneous fishing mortality, annual survival, commercial and angler exploitation, total abundance and catch for each year are presented in Appendix 9. Average parameter values during the early period (1978-82) were: survival, 0.57 ; instantaneous fishing mortality, 0.37 ; commercial exploitation, 0.07 ; and angler exploitation, 0.22 . Average abundance and catch were 12.2 and 3.4 million fish, respectively. Average parameter values during the second period were $0.68,0.19,0.06$, and 0.10 ; while, average abundance and catch were 34.5 and 5.2 million fish. Commercial exploitation has been a relatively small part of the fishery in the western basin of Lake Erie during the entire period. The abundance and catch estimates and the observed catch for the western basin during the entire 10 -year period are given in Figure 13. The CAGEAN results show increased walleye abundance and catch following recruitment of the strong 1982 year class with fairly good agreement between observed and estimated catch.

The CAGEAN program was also used to analyze subsets of data from Lake St. Clair and western Lake Erie to permit generalized comparisons of walleye abundance between the two areas. The period from 1978-85 was selected because catch and effort data from two gear types, angling and trap-net samples, were available for Lake St. Clair during that period. Lake St. Clair anglet harvest data (Table 15) were available from Ontario's waters for the entire period (D. MacLennan, Ontario Ministry of Natural Resources, unpublished data) and from Michigan's waters for 1983 and 1984 only (Haas et al. 1985). We assumed that the relationship between the Ontario and Michigan sport fishery was constant during this 8 -year period. Ontario effort and harvest estimates for all years except 1983 and 1984 were used to estimate Michigan effort and harvest with the ratio of the Michigan to Ontario effort and harvest in those 2 years. CAGEAN estimates of walleye abundance and harvest for Lake St. Clair and western Lake Erie are shown in Figures 14 and 15. Similar trends in abundance and catch were apparent in the two lakes with more obvious deviations in harvest.

The CAGEAN estimates of abundance and harvest in Lake St. Clair were roughly $10 \%$ of the western Lake Erie vaiues. Lake St. Clair is approximately one third as large as the western basin of Lake Erie. The ratio of lake size to estimated walleye abundance was 3.3 which suggested that walleye density was considerably lower in Lake St. Clair. These results are consistent with the grand mean trap-net CPUE values for walleye which were 22.9 in Lake Erie and 8.5 in Lake St. Clair (ratio $=2.7$ ) (Appendices 4 and 5). The CAGEAN estimates are not independent of the trap-net CPUE values since they were part of the input data; however, the angler effort and catch data (used by the CAGEAN model) were more influential in determining abundance estimates. The average angler CPUE was 0.13 in Lake St. Clair and
0.46 in Lake Erie. The temporal pattern of angler tag reporting shows that Lake St. Clair walleye were exploited over a longer period of the year (Figures 8 and 9). The higher tag reporting rate for Lake St. Clair walleye (Table 7) suggests that exploitation was slightly higher. The Lake St. Clair anglers sought a wider variety of fish species compared to Lake Erie anglers (R. Haas, personal observation); and the Lake St. Clair fish community, with its greater species evenness, provided a wider variety of species that were abundant enough to support fisheries.

CAGEAN estimates of harvest were compared to the creel survey harvest estimates which were used as observed inputs to the model (Figures 16 and 17). There was good agreement between the observed and estimated values except for the early years in Lake St. Clair for which the CAGEAN estimates were substantially lower than the creel estimates. We have no explanation for this.

## DISCUSSION

## Net samples, growth, and abundance

The comparison of relative year-class strengths as measured by the trap-net and gill-net samples in Lake Erie indicates that these gear types provide similar information. Trends or changes in the relative abundance of year classes are apparent from examinations of the CPUE data from either sample.

It is apparent from the growth analyses that growth regimes differ in Lake Erie and Lake St. Clair. In general, mean total length-at-age is greater for Lake Erie than for Lake St. Clair walleye. In addition, a change in growth may have occurred in Lake Erie walleye (1978-82 versus 1983-87)

If there is a relationship between year-class abundance and year-class relative growth, it appears to be stronger among Lake Erie walleye than among Lake St. Clair walleye. Year-class specific growth variations in Lake St. Clair walleye are not linearly related to year-class abundance, whereas growth variations in Lake Erie walleye are. Variations in relative growth in Lake St. Clair walleye seem to be influenced by the relative abundance of two or three older age classes.

These observed differences in relative growth and abundance may be examined in light of feeding studies. The abundance of walleye is greater in Lake Erie (higher values of CPUE) where competition for food resources may be high. Age- 0 walleye from western Lake Erie prey mainly on shiners (Notropis atherinoides and Notropis hudsonius) and young-of-the-year clupeids (Dorosoma cepedianum and Alosa pseudoharengus) (Knight et al. 1984). Thus, fluctuations in recruitment of prey may affect year-class strength of walleye in Lake Erie (Knight et al. 1984). No diet studies are available for Lake St. Clair walleye.

The high Shannon-Weiner index for Lake St. Clair suggests that individual prey fishes may be spread more evenly among species so that alternate prey species may be more available. Invertebrates in Lake St. Clair may provide an additional source of food items and may broaden the food base, thereby relaxing competition for fish species. For example, Hexagenia, which are abundant in Lake St. Clair, but rarely found in Lake Erie, may constitute a significant portion of the diet of Lake St. Clair walleye. A broad food base which includes invertebrates may provide a more stable resource when recruitment variability of forage fish is high.

Population modeling of the Lake Erie walleye in the western basin using the CAGEAN model suggests that average abundance during the past 10 years has been about 23 million fish with an upward trend: walleye abundance may have been about 40 million fish since 1984. Hernandez (1988) modeled the western basin fish community and predicted that walleye abundance would eventually reach equilibrium at about 40 million fish under predator-prey conditions that existed in the western basin of Lake Erie during the 1970's and early 1980's.

## Tag-recapture study

Extensive tag recovery data demonstrated that the Lake St. Clair and Lake Erie walleye have a much stronger tendency to move upstream than downstream following spawning activity. Apparently, very few Lake St. Clair walleye migrate to Lake Erie. The temporal pattern of Lake St. Clair tag recoveries from Lake Erie indicated that many were probably Lake Erie fish that had already migrated to Lake St. Clair before capture and tagging. However, there was substantial migration of Lake Erie walleye into Lake St. Clair and the St. Clair and Detroit rivers where they contributed to the angler harvest. Angler recoveries during subsequent spawning seasons indicate that the Thames River in Lake St. Clair and the Maumee River in Lake Erie provide spawning habitat for the majority of walleye present at our tag sites during spring. Analyses of tag reports from the upper half of the St. Clair River showed that $23 \%$ were from Lake Erie. We cannot determine the reliability (e.g., with confidence limits) of the stock composition estimates derived from tag recovery ratios and we caution their strict interpretation. However, it should be noted that although the estimate for Lake St. Clair ( $8 \%$ Lake Erie walleye) is lower than the reported combined estimate from genetic analysis ( $23 \%$ ), it equals the estimate from the same analysis for the sample collected in November 1983. Tag recovery ratios provide an additional set of stock composition estimators. Almost all of the walleye harvested in Lake Erie (95\%) and the majority (70\%) harvested in the Detroit River come from the Lake Erie population. The Lake St. Clair stock contributes greatly to the angler fishery in Lake St. Clair and the northern section of the St. Clair River (approximately $92 \%$ and $77 \%$ of the walleye harvested in these areas were Lake St. Clair fish). The Lake St. Clair stock contributed a modest percentage (30\%) to the walleye harvest in the Detroit River. These
estimates in conjunction with estimates of the average annual harvest may be used to examine the pattern of exploitation for a particular stock. For example, an average of approximately 263,000 Lake St. Clair walleye were harvested annually from the Lake St. Clair-Lake Erie system by anglers in recent years (see Appendix 7). Forty-six percent were taken from Lake St. Clair, $35 \%$ from the northern St. Clair River, and $19 \%$ from the Detroit River. This suggests that management measures for the Lake St. Clair stock must not be geographically confined to Lake St. Clair; substantial harvest of this stock occurs in the northern St. Clair River. A similar analysis of the exploitation pattern for the Lake Erie stock indicates that the harvest of this stock outside of Lake Erie is negligible, on the order of 150,000 fish compared to 7 million in Lake Erie proper.

Reward tags ranging from $\$ 2.00$ to $\$ 8.00$ provide similar return rates from the fishery and these return rates are higher than those from standard tags. An increase in the size of the reward (say, $\$ 100.00$ ) may increase the number of tags reported and permit a less biased estimate of non-reporting. The reward should be of sufficient amount to approach $100 \%$ reporting; i.e., as long as the reward tag is seen, it is a good assumption that it will be reported. Paulik (1963) cautions that a change in the value of the reward may not be the sole factor to account for an increase in reporting rate. For example, we noted an increase in the percent of tags returned during the first year from 1981-83 for both reward and non-reward tags. A follow-up reward tag study should parallel the first study (1981-83) as closely as possible to ensure that targeted fish populations and patterns of exploitation remain similar.

The recovery rates for reward tags were highly variable and similar to recovery rates of standard tags in the third, fourth, and fifth years. Chadwick (1968) found similar results in a 4 -year study of the California striped bass fishery. After the first year, differences in returns from reward tags and standard tags could not be attributed to a true difference in nonresponse (Chadwick 1968). Recovery rates for reward-tagged fish and standard-tagged fish are low beginning in the second year; reward tag returns ranged from 0 to 5 fish (Chadwick 1968). Thus, sample size is an important factor in determining the observed variability in recovery rates in later years. Studies of long duration of populations with high average survival rates are prone to negatively biased estimates of recovery rates in the middle of the study if tagged individuals have varying survival and recovery rates (Pollock and Raveling 1982).

The estimated non-reporting rate for walleye, $33 \%$, seems average relative to those for other species (Table 9). Because the return incentives were low, we feel that the assumption of $100 \%$ reporting of reward tags is unjustified and partitioning of total mortality rate into fishing and natural mortality components is unwarranted. Green et al. (1983) caution that fishing mortality rates are underestimated if tag recovery data are adjusted with a biased reporting rate ratio. Although we cannot be sure of $100 \%$ reporting, the reward tag study does indicate that a
monetary reward improves reporting rate. Even a small amount can improve reporting rate by approximately one and one-half times.

## Genetic analysis

The lack of significant genetic structuring within Lake Erie populations suggests that considerable gene flow does occur. The absence of significant genetic heterogeneity between samples from the Clinton and Thames rivers on Lake St. Clair supports the conclusion that these two localities represent the same gene pool. In fact, a number of walleye tagged during the spawning season in the Clinton River were recovered during subsequent spawning runs in the Thames River (R. Haas, Michigan Department of Natural Resources, unpublished data). However, there is evidence among the Lake Erie samples that the western basin fish do not represent a single, panmictic population. It is possible that the period of low walleye abundance in Lake Erie, during the late 1950's to early 1960's, encouraged interbreeding by otherwise discrete spawning populations.

The genetic differentiation of Lake St. Clair walleye from those of Lake Erie suggest that fry hatched in a spawning area return to that same area to spawn as adults. This conclusion is consistent with the results of this study because the observed genetic structure would not be maintained if young, maturing walleye chose spawning sites at random. Analyses of fall samples from Lake St. Clair corroborated the results of the tag study that significant numbers of Lake Erie walleye mix with the Lake St. Clair population during non-spawning times.

The ability to examine the mitochondrial genome (i.e., mtDNA) in fish has provided researchers with another tool potentially useful for stock delineation (Chapman and Powers 1984; Avise 1987). Mitochondrial DNA is not inherited according to principles of Mendelian genetics; rather, it is transferred from the female parent to her offspring (see e.g., Avise and Lansman 1983). Thus, the term clone has been used to identify unique mtDNA polymorphisms among groups of related conspecifics. This technique, recently applied to Great Lakes walleye populations, supported the existence of two groups of walleye (Billington and Hebert 1988). One group predominates in the eastern Great Lakes (Lake Ontario, Lake Simcoe) and was postulated to be derived from colonization from the Atlantic refugium. Another group predominated in the west (Muskegon, Tittabawassee, and Thames rivers, Lake St. Clair, and western Lake Erie) and was associated with the Mississippi refugium (Billington and Hebert 1988). The greatest clonal diversity was found among walleye from Lake St. Clair and western Lake Erie. However, no particular clones could be associated with individual spawning sites, i.e., no further stock delineation was supported by these data. These results are coincident with our observation that spawning groups associated with any particular river or reef (in Lake Erie) appear to interbreed with groups spawning at adjacent sites. Hence, we
could delineate only two major groups. In addition, both studies report a high level of polymorphisms among walleye from Lake Erie. Despite these similarities, Billington and Hebert (1988) could not separate groups from Lake St. Clair and Lake Erie. We suggest that part of the difficulty may have been due to the nature of their samples; approximately $18 \%$ of the Lake St. Clair and $23 \%$ of the Lake Erie walleye examined were collected in October, a time we found characterized by mixing of stocks.

## RECOMMENDATIONS

Walleye management in Lake St. Clair should not necessarily proceed from the same premises as management in Lake Erie because the two lakes have demonstrably different populations. Management of the Lake Erie and Lake St. Clair populations are complicated by the probable large-scale movement of Lake Erie walleye into Lake St. Clair. Haas et al. (1985) estimated that the average annual sportfishing effort in 1983 and 1984 in Michigan's waters of the St. Clair River was 810,000 angler hours; Lake St. Clair, 2,000,000 angler hours; and in waters of the Detroit River, $1,409,000$ angler hours. If we assume that angling effort in the Ontario waters of these two rivers occurs at a similar level, then total fishing effort in the Lake St. Clair system exceeds angling effort in the western basin of Lake Erie ( $5,600,000$ angler hours; Walleye Task Group, Great Lakes Fishery Commission). We recommend that effective angler effort for walleye harvest be measured every year for Lake St. Clair and the St. Clair and Detroit rivers, as well as for western Lake Erie. Knowledge of angler exploitation rates from southern Lake Huron to western Lake Erie will be a key to effective management of walleye stocks. Although Lake Erie supports a greater population of walleye, exploitation is concentrated in other areas. The fishing pressure on individual stocks cannot be determined without an accurate assessment of stock contribution and angler exploitation rate by area.

We also recommend reward-tag studies on Lake Erie and Lake St. Clair walleye, using large enough reward denominations ( $\$ 100.00$ ), so that walleye exploitation can be measured with acceptable accuracy. Estimates from the catch-at-age (CAGEAN) model indicate that angler exploitation rates are approximately three times commercial exploitation rates for western Lake Erie walleye (range: 1 to 7 times). Because the angler harvest is a major portion of the total harvest, estimates of exploitation rates are necessary to calculate fishing mortality rates. Walleye managers need improved, consistent measures of sportfishing effort and catch in the St. Clair-Detroit River system; they need to know the level of walleye exploitation in Lake St. Clair and Lake Erie; and they need to understand the impact of St. Clair-Detroit River system harvests of walleye originating from Lake Erie populations. Finally, we recommend that walleye diet studies be conducted on Lake St. Clair, presumably a richer food community, where growth rate and relative abundance do not appear to be causally related.

## ACKNOWLEDGMENTS

The work involved in this study was completed by the combined effort of several fisheries agencies. We especially thank K. Muth, U. S. Fish and Wildlife Service, Sandusky, Ohio, for doing the 1982 and 1983 fall Lake Erie gill netting when we were busy with other commitments, and to D. MacLennan, Ontario Ministry of Natural Resources, Tilbury, Ontario, who supplied creel survey data for the Canadian waters of Lake St. Clair. Thanks to all members of the Walleye Task Group, Lake Erie Committee, Great Lakes Fishery Commission, for supplying walleye fishery data. We also thank the following people and their agencies for provision of invaluable samples of spawning walleye: D. Davies, Ohio Department of Natural Resources, Sandusky; D. Allison, Ohio Department of Natural Resources, Findlay; K. Muth, U. S. Fish and Wildlife Service, Sandusky, Ohio; S. Nepszy, Ontario Ministry of Natural Resources, Wheatiey; D. MacLennan, Ontario Ministry of Natural Resources, Tilbury; R. Spitler, Michigan Department of Natural Resources, Pontiac; and R. Lange, New York Department of Environmental Conservation, Oswego. The following MDNR employees were instrumental in data collection and tabulation: biologist, W. Bryant; boat captain, L. Shubel; technicians, G. Olsen, J. Hodge, K. Koster, and J. Clevenger. Thanks also to G. Zurek for manuscript preparation and to supervisor W. C. Latta for editing.

## SPRING SPAWNING SAMPLES

| (1) Clinton River | (4) Sandusky River |
| :--- | :--- |
| (2) Thames River | (5) Niagara Reef |
| (3) Maumee River | (6) Kelleys Island |

(7) Pt. Pelee

FALL MIXED-STOCK SAMPLES
(8) Western Basin - Lake Erie
(9) Anchor Bay - Lake St. Clair
(10) Monroe - Lake Erie (Preliminary)


## TRAP-NET SITES

(1) Lake St. Clair
(2) Lake Erie

GILL-NET SITES
A Lake Erie



Figure 2. Relationship between relative error (bias) of the exploitation rate, $\mathrm{E}(\lambda)$, and the reporting rate, $\alpha$, of reward tags. The line is defined by $\mathrm{E}(\lambda)=(1 / \alpha)-1$ (Conroy and Williams 1981).


Figure 3. Mean trap-net CPUE for 20 species and all species combined in Lake St. Clair and Lake Erie.


Figure 4. Von Bertalanffy growth model for Lake St. Clair (dashed line) and Lake Erie (solid line) walleye. Mean length-at-age values for Lake St. Clair (*) walleye and Lake Erie (0) walleye were estimated from the 1972-85 and 1978-85 trap-net samples.


Figure 5. Relationship between relative growth index of Lake Erie walleye and rank transformation of the mean catch per unit effort (CPUE rank) for the 1972-84 year classes. Each point on the graph represents one year class. See text for method of calculation of the relative growth index; CPUE data are from trap-net samples.


Figure 6. Relationship between relative growth index of Lake St. Clair walleye and rank transformation of the mean catch per unit effort (CPUE rank) for the 1964-82 year classes. Each point on the graph represents one year class. See text for method of calculation of the relative growth index; CPUE data are from trap-net samples.


Figure 7. Percent recapture of walleye tagged in Lake St. Clair and Lake Erie reported for five areas.


Figure 8. Monthly percent of reported recaptures of walleye tagged in Lake St. Clair and Lake Erie recaptured in Lake St. Clair.


Figure 9. Monthly percent of reported recaptures of walleye tagged in Lake St. Clair and Lake Erie recaptured in Lake Erie.


Figure 10. Monthly percent of reported recaptures of walleye tagged in Lake St. Clair and Lake Erie recaptured in the north half of the St. Clair River.


Figure 11. Distribution of angler recaptures of Lake St. Clair walleye with reward and non-reward tags during five subsequent fishing seasons.


Figure 12. Dendrogram based on unweighted pair-group cluster analysis of Nei's (1978) unbiased minimum genetic distance. Significant separation based on genetic distance occurs at the level separating Lake St. Clair and Lake Erie populations.


Figure 13. Estimates of walleye abundance and catch from the CAGEAN analysis and observed walleye catch in the western basin of Lake Erie.


Figure 14. Estimates of walleye abundance in Lake St. Clair and Lake Erie from CAGEAN analyses of comparable data sets.


Figure 15. Estimates of walleye harvest in Lake St. Clair and Lake Erie from CAGEAN analyses of comparable data sets.


Figure 16. Estimate of walleye harvest from the CAGEAN model compared with observed harvest from the creel surveys for Lake St. Clair.


Figure 17. Estimate of walleye harvest from the CAGEAN model compared with observed harvest from the creel surveys for western Lake Erie.

Table 1. Estimated mean catch per unit effort (CPUE) for the 1977-82 year classes of Lake Erie walleye taken in gill nets and trap nets. Only ages $1-5$ are represented in these means.

| Year class | Gill net | Trap net |
| :---: | :---: | :---: |
| 1977 | 32.9 | 7.0 |
| 1978 | 11.6 | 1.6 |
| 1979 | 13.8 | 1.5 |
| 1980 | 17.9 | 4.0 |
| 1981 | 13.8 | 3.0 |
| 1982 | 57.9 | 18.0 |

Table 2. Mean length (mm) and one standard deviation (in parentheses) for 2 to 10 yearold walleye for Lake St. Clair and Lake Erie. The $t$-statistic for comparison of age-specific means is presented; asterisks denote significance at $P \leq 0.05$.

|  | Age |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Location <br> and test | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |
| Lake St. Clair | 339.9 | 385.3 | 430.7 | 480.3 | 512.6 | 546.4 | 577.7 | 617.4 | 625.5 |  |
|  | $(23.71)$ | $(30.93)$ | $(37.38)$ | $(40.69)$ | $(45.36)$ | $(39.26)$ | $(45.46)$ | $(47.85)$ | $(51.02)$ |  |
| Lake Erie | 337.4 398.1 451.1 490.2 527.4 553.0 596.4 631.0 608.5  <br>  $(22.86)$ $(29.43)$ $(35.79)$ $(40.62)$ $(44.68)$ $(47.23)$ $(60.16)$ $(48.98)$ $(109.02)$ <br> $\boldsymbol{t}$-statistic $-3.58^{*}$ $16.52^{*}$ $16.07^{*}$ $5.17^{*}$ $4.41^{*}$ 1.58 $2.85^{*}$ 1.24 -0.68 |  |  |  |  |  |  |  |  |  |

Table 3. Mean total length-at-age (mm) of walleye year classes from Lake Erie trap nets and relative growth measures for year classes. See text for def inition of relative growth.

| Year <br> class | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | Age <br> Relative <br> growth |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1971 | - | - | - | - | - | - | - | 670 | - | - |
| 1972 | - | - | - | - | - | 551 | 665 | 692 | - | 189.35 |
| 1973 | - | - | - | - | 502 | 572 | 547 | 642 | 593 | 9.73 |
| 1974 | - | - | - | 460 | 521 | 521 | 553 | 599 | 695 | 47.63 |
| 1975 | - | - | 417 | 462 | 477 | 517 | 546 | 614 | 649 | -30.78 |
| 1976 | - | 370 | 447 | 443 | 473 | 519 | 553 | 621 | 655 | 27.86 |
| 1977 | 248 | 350 | 397 | 446 | 472 | 517 | 544 | 609 | 590 | -115.51 |
| 1978 | - | 331 | 405 | 453 | 492 | 517 | 535 | 579 | 620 | -120.67 |
| 1979 | 225 | 343 | 413 | 475 | 507 | 539 | 551 | 582 | - | -20.65 |
| 1980 | 228 | 339 | 404 | 463 | 498 | 544 | 571 | - | - | 15.18 |
| 1981 | 219 | 340 | 405 | 457 | 507 | 530 | - | - | - | -11.80 |
| 1982 | 256 | 325 | 387 | 446 | 490 | - | - | - | - | -32.72 |
| 1983 | - | 328 | 399 | 458 | - | - | - | - | - | -21.43 |
| 1984 | 212 | 339 | 413 | - | - | - | - | - | - | -22.50 |
| 1985 | 255 | 351 | - | - | - | - | - | - | - | 28.74 |
| 1986 | 247 | - | - | - | - | - | - | - | - | - |
| Mean | 236 | 342 | 409 | 456 | 494 | 533 | 563 | 623 | 634 | - |

Table 4. Mean total length-at-age (mm) of walleye year classes from Lake St. Clair trap nets and relative growth measures for year classes. See text for definition of relative growth.

| $\begin{aligned} & \text { Year } \\ & \text { class } \end{aligned}$ | Age |  |  |  |  |  |  |  |  |  |  | Relative growth |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |  |
| 1962 | - | - |  | - | - | - | - | - | 655 | 669 | - | - |
| 1963 | - | - | - | - | - | - | - | 604 | 627 | 612 | 703 | - |
| 1964 | - | - | - | - | - | - | 564 | 601 | 597 | 717 | - | -20.3 |
| 1965 | -- | - | - | - | - | 551 | 580 | 592 | 619 | 775 | - | 68.7 |
| 1966 | - | - | - | - | 507 | 533 | 551 | 591 | 668 | 682 | - | -33.4 |
| 1967 | - | - | - | 497 | 517 | 541 | 550 | 615 | - | 669 | - | -25.1 |
| 1968 | - | - | 465 | 511 | 514 | 557 | 600 | 639 | 627 | 694 | - | 108.0 |
| 1969 | - | 423 | 461 | 489 | 539 | 580 | 615 | 664 | 659 | - | - | 206.3 |
| 1970 | 355 | 385 | 436 | 494 | 530 | 547 | 573 | 630 | 652 | 647 | 636 | -35.5 |
| 1971 | 355 | 399 | 430 | 496 | 525 | 557 | 574 | 611 | 612 | 637 | 691 | -33.6 |
| 1972 | 336 | 372 | 463 | 485 | 514 | 541 | 580 | 609 | 626 | 635 | 671 | -87.3 |
| 1973 | 351 | 427 | 439 | 466 | 519 | 553 | 578 | 575 | 631 | - | - | -31.8 |
| 1974 | 361 | 402 | 434 | 495 | 544 | 528 | 558 | 595 | 618 | - | - | -34.3 |
| 1975 | 359 | 409 | 465 | 512 | 516 | 545 | 604 | 679 | - | - | - | 150.7 |
| 1976 | 361 | 434 | 468 | 449 | 505 | 543 | 581 | 581 |  | - | - | -16.6 |
| 1977 | 345 | 380 | 415 | 467 | 501 | 538 | 621 | - | - | - | - | -58.1 |
| 1978 | 339 | 400 | 441 | 456 | 505 | 540 | - | - | - | - | - | -61.8 |
| 1979 | 348 | 393 | 431 | 479 | 531 | - | - | - | - | - | - | -16.0 |
| 1980 | 342 | 379 | 445 | 500 | - | - | - | - | - | - | - | -13.0 |
| 1981 | 343 | 403 | 455 | - | - | - | - | - | - | - | - | 8.7 |
| 1982 | 326 | 381 |  |  | - |  | - | - | - |  | - | -39.3 |
| Mean | 348 | 399 | 446 | 486 | 519 | 547 | 581 | 613 | 633 | 674 | 675 |  |

Table 5. Percent of Lake St. Clair ( $N=1,333$ ) and Lake Erie ( $N=1,159$ ) walleye tag returns from five geographic areas.

| Tag <br> recovery <br> area | Tag station |  |
| :---: | :---: | :---: |
|  | Lake St. Clair | Lake Erie |
| Lake Huron | 7.50 | 1.55 |
| St. Clair River | 29.93 | 8.11 |
| Lake St. Clair | 54.16 | 5.18 |
| Detroit River | 5.40 | 14.41 |
| Lake Erie | 3.00 | 70.75 |

Table 6. Comparison of median Julian day of tagging and tag recovery day for Lake Erie and Lake St. Clair walleye.

| Tagged in: | Median tag day recovered in: |  | Median recapture day in: |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Lake St. Clair | Lake Erie | Lake St. Clair | Lake Erie |
| Lake St. Clair | $\begin{gathered} 152 \\ \text { (June 1). } \end{gathered}$ | $\begin{gathered} 151 \\ \text { (May 31) } \end{gathered}$ | $\begin{gathered} 196 \\ \text { (July } 15 \text { ) } \end{gathered}$ | $\begin{gathered} 155 \\ \text { (June 4) } \end{gathered}$ |
| Lake Erie | $\begin{gathered} 115 \\ \text { (April 25) } \end{gathered}$ | $\stackrel{116}{\text { (April 26) }}$ | $\begin{gathered} 188 \\ \text { (July 7) } \end{gathered}$ | $\begin{gathered} 180 \\ \text { (June 29) } \end{gathered}$ |

Table 7. Mean annual survival rate, recovery rate, and adult life span (years) from walleye tag analyses using the BROWNIE model. These estimates are annual means derived from a tagging experiment began in 1975 in Lake St. Clair and 1978 in Lake Erie and include tag recoveries through 1987. SE is standard error of the mean.

| Tag site | Mean <br> survival <br> rate | Mean <br> recovery <br> rate | Mean <br> adult <br> lifespan |
| :--- | :---: | :---: | :---: |
| Lake St. Clair | 53.28 | 5.32 | 1.59 |
| SE | 1.75 | 0.24 | 0.08 |
| $95 \%$ confidence interval | $(49.85-56.71)$ | $(4.86-5.79)$ | $(1.44-1.76)$ |
| Lake Erie | 60.93 | 3.21 | 2.02 |
| SE | 1.96 | 0.13 | 0.13 |
| $95 \%$ confidence interval | $(57.10-64.77)$ | $(2.95-3.47)$ | $(1.78-2.30)$ |

Table 8. Reporting rate (ratio of reward-tag to standard-tag return rate) for tagging years 1981-83 for Lake St. Clair walleye. $N$ is number of walleye released with reward tags ( $\$ 2.00-\$ 8.00$ ) and $n$ is number of walleye released with standard tags. The year of tagging is indicated by $i$, the year following by $i+1$, etc.

| Year | $N$ | $n$ | i | $\mathrm{i}+1$ | $\mathrm{i}+2$ | $\mathrm{i}+3$ | $\mathrm{i}+4$ | $\mathrm{i}+5$ |
| :---: | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 317 | 650 | 2.237 | 1.212 | 1.709 | 2.051 | 4.101 | 2.051 |
| 1982 | 198 | 259 | 1.744 | 1.047 | 0.727 | 6.540 | 0.436 | - |
| 1983 | 394 | 1,134 | 1.531 | 1.367 | 2.878 | 2.878 | 1.151 | - |

Table 9. Range of non-reporting rates and average non-reporting rates from fish tagging and bird banding studies. Non-reporting rates and their means were taken from cited references or were estimated from data in cited references.

| Species | Area | Year | Non-reporting rate (percent) |  | Reward | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Range | Mean |  |  |
| FISH |  |  |  |  |  |  |
| Rainbow trout | Lake Lanier, GA | 1980-81 | - | 61 | \$5.00 | Weaver \& England 1986 |
| Various estuarine species ${ }^{1}$ | Texas Bay systems | 1976-78 | 64-83 | 71 | $\begin{array}{r} \$ 1.00 \\ \text { to } \\ \$ 25.00 \end{array}$ | Green et al. 1983 |
| Largemouth bass | Centerhill Reservoir, TN | - | - | 46 | - | Coomer 1976 (cited in Green et al. 1983) |
| Largemouth bass | Merle Collins Reservoir, CA | 1966-69 | 15-54 | 34 | $\begin{array}{r} \$ 1.00 \\ \text { and } \\ \$ 5.00 \end{array}$ | Rawstron 1972 |
| Largemouth bass White catfish Bluegill | Folsom Lake, CA | 1962-68 | $\begin{gathered} 5-52 \\ 22-66 \end{gathered}$ | $\begin{aligned} & 38 \\ & 39 \\ & 69 \end{aligned}$ | $\begin{aligned} & \$ 5.00 \\ & \$ 5.00 \\ & \$ 5.00 \end{aligned}$ | Rawstron 1971 |
| Rainbow trout | Lake Tahoe, CA | - | $\begin{gathered} 36 .-39^{2} \\ 51-53^{3} \end{gathered}$ | - | $\begin{array}{r} \$ 1.00 \\ \text { and } \\ \$ 5.00 \end{array}$ | Rawstron, unpubl. (in Rawstron 1971) |
| Striped bass | Sacramento- <br> San Joaquin <br> River system, CA | 1958-61 | 31-47 | 38 | \$5.00 | Chadwick 1968 |
| Cod | Barents Sea (U.K. Fleet) | 1959-61 | 56-67 | 59 | - | Margetts 1963 |
| Cod | West Greenland \& North Atlantic | 1953-59 | 69-83 | 77 | -s | Horsted 1963 |
| Yellowfin \& skipjack tuna | Eastern Pacific | 1955-59 | 10-50 | - | - ${ }^{4}$ | Schaeffer et al. 1963 |
| Rainbow trout | Kern River, CA | 1956-57 | - | 38 | \$5.00 | Butler 1962 |
| Channel catfish | Sacramento Valley rivers, CA | 1955 | $30-40^{6}$ | $40^{6}$ | $\$ 200.00$ <br> Lottery | McCammon \& LaFaunce 1961 |
| Rainbow trout | CA | - | 40-60 | - | -' | Butler, unpubl., cited in McCammon \& LaFaunce 1961 |
| Channel catfish | Colorado River, CA | 1953-54 | - | 15 | Lottery ${ }^{8}$ | McCammon 1956 |
| Brook trout Brown trout | Millers River, MA | $\begin{aligned} & 1953 \\ & 1953 \end{aligned}$ | $\begin{aligned} & 9-25 \\ & 9-25 \end{aligned}$ | 17 25 | -' | Mullan 1959 |
| Rainbow trout |  | 1953 | 9-25 | 17 | -' |  |
| Brook trout <br> Brown trout | Squannacook River, MA | 1953 1953 | - | 60 60 | -' | Mullan 1959 |
| Rainbow trout |  | 1953 | - | 60 | -' |  |

Table 9. Continued:

| Species | Area | Year | Non-reporting rate (percent) |  | Reward | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Range | Mean |  |  |
| Largemouth bass | Lakes, ponds, and | 1950-53 | - | 25 | \$1.00 ${ }^{10}$ | Stroud and Bitzer 1955 |
| Smallmouth bass | reser voirs, MA |  | - | 25 | \$1.00 ${ }^{10}$ |  |
| Brown bullhead |  |  | - | 25 | \$1.00 ${ }^{10}$ |  |
| White perch |  |  | - | 25 | \$1.00 ${ }^{10}$ |  |
| Yellow perch |  |  | - | 25 | \$1.00 ${ }^{\text {20 }}$ |  |
| Chain pickerel |  |  | - | 25 | Lottery |  |
| BIRDS |  |  |  |  |  |  |
| American black duck | Canada \& U. S: <br> Atlantic \& Mississippi flyways | 1978-80 | 0-69 | 57 | \$15.00 | Conroy \& Blandin 1984 |
| Mallard | U. S. \& Canada | 1972-73 | 0-94 | - | \$10.00 | Henny \& Burnham 1976 |
| Mourning dove | U.S. | 1970-72 | 35-69 | 58 | Book | Reeves 1979 |
| Mourning dove | U.S. | 1965-66 | 59-64 | 60 | Book | Tomlinson 1968 |
| Waterfowl (10 species) | U.S. | 1954-58 | 50-57 | 54 | - ${ }^{11}$ | Geis \& Atwood 1961 |
| Mallard | U.S. | 1954-57 | 32-55 | 50 | - ${ }^{11}$ | Geis \& Atwood 1961 |
| Mallard | IL | 1948-51 | 0-65 | 50 | \$2.00 | Bellirose 1955 |
| Mallard | Canada \& U.S: <br> Mississippi flyway | 1950-51 | 38-62 | 56 - | \$2.00 | Bellrose 1955 |

${ }^{1}$ Spotted seatrout, red drum, other seatrout, flounder, Atlantic croaker, black drum, and sheepshead.
${ }^{2}$ First year.
${ }^{3}$ Second year.
${ }^{4}$ Placed tagged fish in catch of commercial vessels.
${ }^{3}$ Assumed one segment of fishery reported all tags.
${ }^{6}$ Used estimates from the literature.
' Estimated non-reporting rates from known harvest rate.
${ }^{2}$ Used two types of tags; estimated non-reporting rate by assuming all recoveries of one type of tag were reported.
${ }^{9}$ Estimated non-reporting rates from partial creel census.
${ }^{10}$ Gift certificate.
${ }^{11}$ Estimated non-reporting rate from follow-up survey questionnaire.

Table 10. Allelic frequencies at polymorphic loci, probability of conformance to HardyWeinberg expectations (in parentheses), and $P$, probability of genetic homogeneity for Lake Erie walleye populations. $N$ is sample size. Single asterisk denotes significance at $P \leq 0.05$.

| Locus | Population |  |  |  |  |  | $P$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pelee | W. Basin | Maumee | Niagara | Kelley | Sandusky |  |
| ADH-1 | (0.209) | (0.396) | (0.781) | (0.435) | (0.763) | (0.401) | 0.075 |
| $N$ | 49 | 72 | 388 | 281 | 119 | 97 |  |
| 75 | 0.224 | 0.160 | 0.170 | 0.189 | 0.172 | 0.144 |  |
| 100 | 0.776 | 0.833 | 0.830 | 0.811 | 0.828 | 0.856 |  |
| 140 | 0.000 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 |  |
| GMP-3 | (0.944) | (0.349) | (0.641) | (0.807) | (0.596) | (0.044)* | 0.561 |
| $N$ | 49 | 72 | 393 | 290 | 120 | 98 |  |
| 75 | 0.449 | 0.389 | 0.438 | 0.398 | 0.404 | 0.388 |  |
| 100 | 0.551 | 0.611 | 0.562 | 0.602 | 0.596 | 0.612 |  |
| IDH-1 | (0.376) | (0.739) | (0.095) | (0.097) | (0.107) | (0.211) | 0.386 |
| $N$ | 49 | 72 | 390 | 288 | 121 | 98 |  |
| 80 | 0.296 | 0.403 | 0.332 | 0.351 | 0.298 | 0.357 |  |
| 100 | 0.704 | 0.597 | 0.667 | 0.649 | 0.702 | 0.638 |  |
| 120 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.005 |  |
| MDH-1 | - | - | (0.980) | - | (0.964) | (0.960) | 0.225 |
| $N$ | 49 | 72 | 393 | 290 | 121 | 98 |  |
| 40 | 0.000 | 0.000 | 0.001 | 0.000 | 0.004 | 0.000 |  |
| 100 | 1.000 | 1.000 | 0.999 | 1.000 | 0.996 | 0.995 |  |
| 155 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 |  |
| MDH-2 | - | - | (0.960) | - | - | - | 0.668 |
| $N$ | 49 | 72 | 393 | 290 | 121 | 98 |  |
| 100 | 1.000 | 1.000 | 0.997 | 1.000 | 1.000 | 1.000 |  |
| 140 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 |  |
| MDH-3 | (0.484) | (0.642) | (0.193) | (0.300) | (0.109) | (0.707) | 0.503 |
| $N$ | 49 | 72 | 393 | 290 | 120 | 98 |  |
| 80 | 0.000 | 0.000 | 0.014 | 0.021 | 0.008 | 0.015 |  |
| 100 | 0.714 | 0.743 | 0.733 | 0.734 | 0.754 | 0.684 |  |
| 120 | 0.286 | 0.257 | 0.253 | 0.245 | 0.237 | 0.301 |  |
| PGM-1 | - | 72 | - | - | (0.963) | (0.960) | - |
| $N$ | 49 | 72 | 393 | 290 | 120 | 98 |  |
| 70 | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 | 0.005 |  |
| 100 | 1.000 | 1.000 | 1.000 | 1.000 | 0.996 | 0.995 |  |
| PGI-1 | - | 7 | - | - | (0.964) | - | 0.188 |
| $N$ | 49 | 72 | 393 | 290 | 121 | 98 |  |
| 100 | 1.000 | 1.000 | 1.000 | 1.000 | 0.996 | 1.000 |  |
| 270 | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 | 0.000 |  |

Table 10. Continued:

|  | Population |  |  |  |  |  |  |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Locus | Pelee | W. Basin | Maumee | Niagara | Kelley | Sandusky | $P$ |
|  |  |  |  |  |  |  |  |
| PGI-2 | - | - | $(0.939)$ | $\overline{29}$ | $(0.964)$ | $\overline{1}$ | 0.552 |
| $N$ | 49 | 72 | 393 | 290 | 121 | 98 |  |
| 100 | 1.000 | 1.000 | 0.996 | 1.000 | 0.996 | 1.000 |  |
| 140 | 0.000 | 0.000 | 0.004 | 0.000 | 0.004 | 0.000 |  |
|  |  | $(0.953)$ | $(0.719)$ | $(0.858)$ |  | $(0.837)$ | 0.168 |
| PGI-3 | 49 | 72 | 393 | 289 | 120 | 98 |  |
| $N$ | 0.000 | 0.007 | 0.018 | 0.010 | 0.000 | 0.020 |  |
| 90 | 0.993 | 0.982 | 0.990 | 1.000 | 0.980 |  |  |
| 100 | 1.000 |  |  |  |  |  | 0.126 |
| Total |  |  |  |  |  |  |  |

Table 11. Allelic frequencies at four polymorphic loci, probability of conformance to Hardy-Weinberg expectations (in parentheses), and probability of genetic homogeneity for Lake Erie and Lake St. Clair walleye. $N$ is sample size. Single asterisk denotes significance at $P \leq 0.05$; double asterisks denote significance at $P \leq 0.01$.

|  | Population |  |  |
| :---: | :---: | :---: | :---: |
| Locus | Lake Erie | Lake St. Clair |  |
|  |  |  |  |
| ADH-1 | $(0.509)$ | $(0.854)$ | 0.141 |
| $N$ | 1,006 | 633 |  |
| 75 | 0.175 | 0.151 |  |
| 100 | 0.825 | 0.849 |  |
| GMP-3 | $(0.447)$ | $(0.828)$ | $0.000^{* * *}$ |
| $N$ | 1,022 | 645 |  |
| 75 | 0.415 | 0.486 |  |
| 100 | 0.585 | 0.514 | $0.018^{*}$ |
|  | $(0.003) * *$ | $(0.475)$ |  |
| IDH-1 | 1,018 | 642 |  |
| $N$ | 0.339 | 0.296 |  |
| 80 | 0.660 | 0.704 | 0.0 |
| 100 | 0.001 | $(0.598)$ |  |
| 120 | $(0.026) *$ | 643 |  |
| MDH-3 | 1,022 | 0.017 |  |
| $N$ | 0.014 | 0.652 |  |
| 80 | 0.731 | 0.331 |  |
| 100 | 0.255 |  | $0.000^{* *}$ |
| 120 |  |  |  |
| Total |  |  |  |

Table 12. Allelic frequencies at polymorphic loci, probability of conformance to Hardy Weinberg expectations (in parentheses), and probability of genetic homogeneity for Lake St. Clair walleye. $N$ is sample size. Double asterisks denote significance at $P \leq 0.01$.

| Locus | Population |  | P |
| :---: | :---: | :---: | :---: |
|  | Clinton River | Thames River |  |
| ADH-1 | (0.144) | (0.256) | 0.700 |
| $N$ | 313 | 320 |  |
| 75 | 0.147 | 0.155 |  |
| 100 | 0.853 | 0.845 |  |
| GMP-3 | (0.566) | (0.792) | 0.975 |
| N | 320 | 325 |  |
| 75 | 0.486 | 0.486 |  |
| 100 | 0.514 | 0.514 |  |
| IDH-1 | (0.456) | (0.077) | 0.751 |
| $N$ | 320 | 322 |  |
| 80 | 0.300 | 0.292 |  |
| 100 | 0.700 | 0.708 |  |
| MDH-2 | - | (0.956) | 0.161 |
| $N$ | 321 | 327 |  |
| 100 | 1.000 | 0.997 |  |
| 140 | 0.000 | 0.003 |  |
| MDH-3 | (0.996) | (0.461) | 0.729 |
| $N$ | 318 | 325 |  |
| 80 | 0.019 | 0.015 |  |
| 100 | 0.659 | 0.645 |  |
| 120 | 0.322 | 0.340 |  |
| PGM-1 | (0.933) | (0.978) | 0.307 |
| $N$ | 321 | 327 |  |
| 70 | 0.005 | 0.002 |  |
| 100 | 0.995 | 0.998 |  |
| PGI-2 | (0.978) | (0.978) | 0.972 |
| $N$ | 321 | 327 |  |
| 100 | 0.998 | 0.998 |  |
| 140 | 0.002 | 0.002 |  |
| PGI-3 | (0.754) | $(0.001) * *$ | 0.846 |
| $N$ | 320 | 316 |  |
| 90 | 0.017 | 0.016 |  |
| 100 | 0.983 | 0.984 |  |
| Total |  |  | 0.916 |

Table 13. Length-at-age ( mm ) for male and female spawning walleye from the Thames and Maumee rivers. $N$ is sample size. $P$ is probability of a significant difference in length-at-age for the two groups; NS is nonsignificant.

| Rivers | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males |  |  |  |  | Females |  |  |  |  |
|  | 3 | 4 | 5 | 6 | 7 | 3 | 4 | 5 | 6 | 7 |
| Thames | 351 | 404 | 461 | 484 | 510 | 367 | 475 | 489 | 516 | 570 |
| $N$ | 33 | 49 | 22 | 17 | 23 | 10 | 11 | 12 | 46 | 32 |
| $P$ | 0.01 | 0.01 | NS | NS | NS | 0.05 | NS | 0.01 | 0.01 | 0.01 |
| Maumee | 390 | 434 | 459 | 491 | 525 | 409 | 490 | 542 | 559 | 611 |
| $N$ | 130 | 75 | 10 | 21 | 7 | 8 | 32 | 7 | 12 | 15 |

Table 14. Somatic weight-at-age (g) for male and female spawning walleye from the Thames and Maumee rivers. N is sample size. P is probability of a significant difference in weight-at-age for the two groups; NS is nonsignificant.

| Rivers | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males |  |  |  |  | Females |  |  |  |  |
|  | 3 | 4 | 5 | 6 | 7 | 3 | 4 | 5 | 6 | 7 |
| Thames | 395 | 554 | 819 | 933 | 1,149 | 438 | 927 | 1,031 | 1,172 | 1,554 |
| $N$ | 33 | 49 | 22 | 17 | 23 | 10 | 11 | 12 | 46 | 32 |
| $P$ | 0.01 | 0.01 | NS | 0.01 | NS | NS | 0.05 | 0.01 | 0.01 | 0.01 |
| Maumee | 526 | 749 | 913 | 1,139 | 1,354 | 609 | 1,099 | 1,524 | 1,716 | 2,134 |
| $N$ | 130 | 75 | 10 | 21 | 7 | 8 | 32 | 7 | 12 | 15 |

Table 15. Estimated angler effort and harvest of walleye from Lake St. Clair

| Year | Ontario fishery |  | Michigan fishery |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Angler hours | Walleye harvest (number) | Angler hours | Walleye harvest (number) |
| 1977 | 82,260 | 11,700 | - | - |
| 1978 | 546,558 | 47,073 | - | - |
| 1979 | 541,909 | 157,602 | - | - |
| 1980 | 580,441 | 133,767 | - | - |
| 1981 | 457,512 | 91,423 | - | - |
| 1982 | 740,684 | 164,021 | - | - |
| 1983 | 459,022 | 85,959 | 1,524,065 | 131,529 |
| 1984 | 486,245 | 119,136 | 1,447,117 | 132,221 |
| 1985 | 270,339 | 82,125 | - | - |
| Mean | 462,774 | 99,201 | 1,485,591 | 131,875 |

## LITERATURE CITED

Allendorf, F. W. and F. M. Utter. 1979. Population genetics of fish. Pages 407-454 in W. S. Hoar, D. S. Randall, and J. R. Brett (eds.). Fish physiology, Volume. 8, Academic Press, New York.

Allendorf, F. W., N. Mitchill, N. Ryman, and G. Stahl. 1977. Isozyme loci in brown trout (Salmo trutta L.): detection and interpretation from population data. Hereditas 86:179-190.

Avise, J. C. 1987. Identification and interpretation of mitochondrial DNA stocks in marine species. Pages 105-136 in H. E. Kumpf (ed), Proceedings of the Stock Identification Workshop, NOAA Technical Memorandum NMFS-SEFC-199.

Avise, J. C. and R. A. Lansman. 1983. Polymorphism of mitochondrial DNA in populations of higher animals. Pages 146-164 in M. Nei and R. K. Koehn (eds.). Evolution of genes and proteins, Sinauer Associates Incorporated, Sunderland, Maryland.

Bannerot, S. P. and C. B. Austin. 1983. Using frequency distributions of catch per unit effort to measure fish stock abundance. Transactions of the American Fisheries Society 112:608-617.

Bellrose, F. C., Jr. 1955. A comparison of recoveries from reward and standard bands. Journal of Wildlife Management 19:71-75.

Billington, N. and P. D. N. Hebert. 1988. Mitochondrial DNA variation in Great Lakes walleye (Stizostedion vitreum) populations. Canadian Journal of Fisheries and Aquatic Sciences 45:643-654.

Brownie, C., D. R. Anderson, K. P. Burnham, and D. S. Robson. 1985. Statistical inference from band recovery data-A handbook (2nd edition). U. S. Department of the Interior, Fish and Wildlife Service, Resource Publication No. 156, Washington, D.C.

Butler, R. L. 1962. Recognition and return of trout tags by California anglers. California Fish and Game 48:5-18.

CAGEAN-PC User Manual. 1987. International Pacific Halibut Commission.
Chadwick, H. K. 1968. Mortality rates in the California striped bass population. California Fish and Game 54:228-246.

Chapman, R. W. and D. A. Powers. 1984. A method for the rapid isolation of mitochondrial DNA from fishes. Maryland Sea Grant Program, Technical Report UM-SG-TS-8405, College Park.

Cochran, W. G. 1977 (3rd edition). Sampling techniques. Wiley and Sons: New York.
Conover, W. J. 1980 (2nd edition). Practical non-parametric statistics. Wiley and Sons, New York.

Conroy, M. J. and W. W. Blandin. 1984. Geographic and temporal differences in band reporting rates for American black ducks. Journal of Wildlife Management 48:2336.

Conroy, M. J. and B. K. Williams. 1981. Sensitivity of band reporting-rate estimates to violations of assumptions. Journal of Wildlife Management 45:789-792.

Craig, J. F. 1980. Sampling with traps. Pages 55-70 in T. Backiel and R. L. Welcomme (eds). Guidelines for sampling fish in inland waters. Food and Agriculture Organization 33, Rome, Italy.

Deriso, R. B., T. J. Quinn II, and P. R. Neal. 1985. Catch-age analysis with auxiliary information. Canadian Journal of Fisheries and Aquatic Sciences 42:815-824.

Direcki, J. A. 1984a. St. Clair River physical and hydraulic characteristics. Great Lakes Environmental Research Laboratory Contribution No. 413, Ann Arbor, Michigan, USA.

Direcki, J. A. 1984b. Detroit River physical and hydraulic characteristics. Great Lakes Environmental Research Laboratory Contribution No. 417, Ann Arbor, Michigan, USA.

Ferguson, R. G., and A. J. Derksen. 1971. Migration of adult and juvenile walleyes (Stizostedion vitreum vitreum) in southern Lake Huron, Lake St. Clair, Lake Erie and connecting waters. Journal of the Fisheries Board of Canada 28:1133-1142.

Geis, A. D. and E. L. Atwood. 1961. Proportion of recovered waterfowl bands reported. Journal of Wildlife Management 25:154-159.

Green, A. W., G. C. Matlock, and J. E. Weaver. 1983. A method for directly estimating the tag-reporting rate of anglers. Transactions of the American Fisheries Society 112:412-415.

Haas, R. C., W. C. Bryant, K. D. Smith, and A. J. Nuhfer. 1985. Movement and harvest of fish in Lake St. Clair, St. Clair River, and Detroit River. Final Report, Winter Navigation Study, Fisheries Division, Michigan Department of Natural Resources, Lansing.

Harris, H. and D. A. Hopkinson. 1977. Handbook of enzyme electrophoresis in human genetics. American Elsevier Publishing Company, New York.

Henny, C. J. and K. P. Burnham. 1976. A reward band study of mallards to estimate band reporting rates. Journal of Wildlife Management 40:1-14.

Hernandez, A. B. L. 1988. Comparative study of the response of simulated temperate and tropical fish communities to fishery exploitation and environmental variability. Ph.D. dissertation, Biology Program, Case Western Reserve University, Cleveland, Ohio.

Holt, S. J. 1963. Tagging experiments and the theory of advertising. Pages $29-30$ in International Commission for the Northwest Atlantic Fisheries, Special Publication No. 4: North Atlantic Fish Marking Symposium. ICNAF Headquarters, Dartmouth, Nova Scotia, Canada.

Horsted, S. A. 1963. On non-reported recaptures from Danish tagging experiments on cod, Subarea 1. Pages 22-25 in International Commission for the Northwest Atlantic Fisheries, Special Publication No. 4: North Atlantic Fish Marking Symposium. ICNAF Headquarters, Dartmouth, Nova Scotia, Canada.

Ihssen, P. E., H. E. Booke, J. M. Casselman, J. M. McGlade, N. R. Payne, and F. M. Utter. 1981. Stock identification: materials and methods. Canadian Journal of Fisheries and Aquatic Sciences 38:1838-1855.

Johnston, D. A. 1977. Population dynamics of walleye (Stizostedion vitreum vitreum) and yellow perch (Perca flavescens) in Lake St. Clair, especially during 1970-76. Journal of the Fisheries Research Board of Canada 34:1869-1877.

Knight, R. L., F. J. Margraf, and R. F. Carline. 1984. Piscivory by walleyes and yellow perch in western iake Erie. Transactions of the American Fisheries Society 113:677-693.

Krementz, D. G., M. J. Conroy, J. E. Hines, and H. F. Percival. 1987. Sources of variation in survival and recovery rates of American black ducks. Journal of Wildlife Management 51:689-700.

Margetts, A. R. 1963. Measurement of the efficiency of recovery and reporting of tags from recaptured fish. Pages 255-257 in International Commission for the Northwest Atlantic Fisheries, Special Publication No. 4: North Atlantic Fish Marking Symposium. ICNAF Headquarters, Dartmouth, Nova Scotia, Canada.

McCammon, G. W. 1956. A tagging experiment with channel catfish (Ictalurus punciatuis) in the lower Colorado River. California Fish and Game 42:323-335.

McCammon, G. W. and D. A. LaFaunce. 1961. Mortality rates and movement in the channel catfish population of the Sacramento Valley. California Fish and Game 47:5-26.

Milner, G. B., D. J. Teel, and F. M. Utter. 1983. Genetic stock identification study. Report, Coastal Zone and Estuarine Studies Division, Northwest and Alaska Fisheries Center, National Marine Fisheries Service, Seattle, Washington.

Mullan, J. W. 1959. Probable effectiveness of tag reporting from trout stocked in two Massachusetts stream drainages. Mimeo. Massachusetts Division of Fisheries and Game.

Nei, M. 1978. Molecular population genetics and evolution. American Elsevier Publishing Company, New York.

Nepszy, S. J. 1977. Changes in percid populations and species interactions in Lake Erie. Journal of the Fisheries Research Board of Canada 34:1861-1868.

Paulik, G. J. 1963. Detection of incomplete reporting of tags. Pages 238-247 in International Commission for the Northwest Atlantic Fisheries, Special Publication No. 4: North Atlantic Fish Marking Symposium. ICNAF Headquarters, Dartmouth, Nova Scotia, Canada.

Pollock, K. H. and D. G. Raveling. 1982. Assumptions of modern band-recovery models with emphasis on heterogeneous survival rates. Journal of Wildlife Management 46:88-98.

Prager, M. H., S. B. Saila, and C. W. Recksiek. 1987. FISHPARM: A microcomputer program for parameter estimation of nonlinear models in fishery science. Old Dominion University, Department of Oceanography, Technical Report 87-10. Norfolk, Virginia.

Rawstron, R. R. 1971. Nonreporting of tagged white catfish, largemouth bass, and bluegills by anglers at Folsom Lake, California. California Fish and Game 57:246-252.

Rawstron, R. R. 1972. Nonreporting of tagged largemouth bass, 1966-1969. California Fish and Gamc 58:145-147.

Reeves, H. M. 1979. Estimates of reporting rates for mourning dove bands. Journal of Wildlife Management 43:36-41.

Regier, H. A., V. C. Applegate, and R. A. Ryder. 1969. The ecology and management of the walleye in western Lake Erie. Great Lakes Fishery Commission Technical Report 15, Ann Arbor, Michigan, USA.

Ricker, W. E. 1975. Computation and interpretation of biological statistics of fish populations. Bulletin of the Fisheries Research Board of Canada 191.

Schaefer, M. B., B. M. Chatwin, and G. C. Broadhead, 1963. Some results of tagging and recovery of tropical tunas 1955-1959. Pages 306-307 in International Commission for the Northwest Atlantic Fisheries, Special Publication No. 4: North Atlantic Fish Marking Symposium. ICNAF Headquarters, Dartmouth, Nova Scotia, Canada.

Schneider, J. C., and J. H. Leach. 1977. Walleye (Stizostedion vitreum vitreum) fluctuations in the Great Lakes and possible causes, 1800-1975. Journal of the Fisheries Research Board of Canada 34:1878-1889.

Sokal, R. R. and F. J. Rohlf. 1981 (2nd edition). Biometry: The principles and practice of statistics in biological research. W. H. Freeman and Company, New York.

Stroud, R. H. and H. Bitzer. 1955. Harvests and management of warm-water fish populations in Massachusetts' lakes, ponds, and reservoirs. Progressive Fish-Culturist 17:51-63.

Swofford, D. L. and R. B. Selander. 1981. BIOSYS-1: a FORTRAN program from comprehensive analysis of electrophoretic data in population genetics and systematics. The Journal of Heredity 72:281-283.

Todd, T. N. 1983. Use of microwave oven simplifies preparation of agar overlays. Isozyme Bulletin 16:92.

Tomlinson, R. E. 1968. Reward banding to determine reporting rate of recovered mourning dove bands. Journal of Wildlife Management 32:6-11.

Vandermeer, J. 1981. Elementary mathematical ecology. John Wiley and Sons, New York.
VanVooren, A. R. 1978. Characteristics of walleye spawning stocks. Ohio Department of Natural Resources, Performance Report, Dingell-Johnson Project F-35-R, Study 4, Columbus.

Vaughan, D. S. and P. Kanciruk. 1982. An empirical comparison of estimation procedures for the von Bertalanffy growth equation. Journal du Conseil International pour l'Exploration de la Mer 40:211-219.

Weaver, O. R. and R. H. England. 1986. Return of tags with different rewards in Lake Lanier, Georgia. North American Journal of Fisheries Management 6:132-133.

Willis, D. W. 1987. Use of gill-net data to provide a recruitment index for walleyes. North American Journal of Fisheries Management 7:591-592.

Wolfert, D. R. 1963. The movement of walleyes tagged as yearlings in Lake Erie. Transactions of the American Fisheries Society 92:414-420.

Yeh, C. F. 1977. Relative selectivity of fishing gear used in a large reservoir in Texas. Transactions of the American Fisheries Society 106:309-313.

Report approved by W. C. Latta

Appendix 1. List of fishes caught during spring trap-net surveys in Lake St. Clair and Lake Erie 1975-87.

| Species | Scientific name | Lake St. Clair | Lake Erie |
| :--- | :--- | :--- | :--- |
| Lake sturgeon | Acipenser fulvescens | Present | Absent |
| Longnose gar | Lepisosteus osseus | Present | Present |
| Bowfin | Amia calva | Present | Present |
| Alewife | Alosa pseudoharengus | Present | Present |
| Gizzard shad | Dorosoma cepedianum | Present | Present |
| Mooneye | Hiodon tergisus | Present | Absent |
| Northern pike | Esox lucius | Present | Present |
| Muskellunge | Esox masquinongy | Present | Present |
| Black bullhead | Ictalurus melas | Present | Present |
| Brown bullhead | Ictalurus nebulosus | Present | Present |
| Channel catfish | Ictalurus punctatus | Present | Present |
| Stonecat | Noturus flavus | Absent | Present |
| American eel | Anguilla rostrata | Present | Absent |
| Burbot | Lota lota | Absent | Present |
| White perch | Morone americana | Present | Present |
| White bass | Morone chrysops | Present | Present |
| Freshwater drum | Aplodinotus grunniens | Present | Present |
| Lake whitefish | Coregonus clupeaformis | Absent | Present |
| Lake trout | Salvelinus namaycush | Present | Absent |
| Coho salmon | Oncorhynchus kisutch | Absent | Present |
| Splake | Salvelinus fontinalis | Prent | Absent |
|  | X S. namaycush | Present | Presh |

Appendix 2. Mean number of fish caught per trap-net lift in Lake St. Clair. Catch per effort is number caught in 24 hours.

| Species | Year |  |  |  |  |  |  |  |  |  |  | Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |  |
| Smallmouth bass | 6.7 | 3.6 | 12.6 | 21.2 | 16.7 | 39.9 | 36.2 | 11.2 | 14.9 | 26.2 | 29.1 | 19.85 |
| Yellow perch | 6.8 | 0.4 | 2.6 | 5.1 | 7.3 | 4.2 | 0.9 | 1.4 | 5.3 | 21.7 | 15.4 | 6.46 |
| Rock bass | 52.5 | 11.8 | 34.5 | 27.8 | 52.8 | 50.1 | 32.4 | 50.0 | 62.5 | 48.8 | 38.5 | 41.97 |
| White bass | 0.7 | 0.1 | 0.9 | 1.1 | 3.9 | 4.3 | 1.6 | 7.7 | 2.8 | 1.4 | 0.7 | 2.29 |
| White perch | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | $<0.1$ | $<0.1$ | <0.1 | <0.1 | 0.01 |
| Pumpkinseed | 3.7 | 8.8 | 11.7 | 4.9 | 5.5 | 25.8 | 12.0 | 29.1 | 10.9 | 27.1 | 51.2 | 17.34 |
| Bluegill | 0.4 | 0.5 | 0.3 | 0.3 | 0.1 | 0.2 | 0.2 | 0.1 | 0.0 | 0.2 | 0.8 | 0.27 |
| Black crappie | 0.6 | 1.5 | 0.4 | 1.0 | 0.6 | 4.2 | 12.0 | 1.6 | 1.5 | 5.6 | 6.8 | 3.25 |
| Channel catfish | 17.1 | 10.5 | 15.2 | 12.7 | 29.6 | 14.9 | 4.9 | 8.0 | 11.9 | 17.3 | 6.1 | 13.47 |
| Brown bullhead | 1.8 | 0.7 | 0.6 | 1.2 | 1.1 | 0.7 | 0.5 | 0.3 | 0.2 | 0.2 | 0.1 | 0.68 |
| Muskellunge | 0.7 | 0.0 | 0.1 | 0.4 | 0.3 | 0.0 | 0.1 | 0.1 | 0.3 | 0.1 | 0.1 | 0.21 |
| White sucker | 1.0 | 0.4 | 0.6 | 1.0 | 0.7 | 0.7 | 0.4 | 1.1 | 0.5 | 0.5 | 0.3 | 0.66 |
| Redhorse | 2.5 | 0.9 | 1.0 | 1.9 | 1.7 | 1.9 | 1.5 | 2.0 | 1.3 | 0.0 | 0.0 | 1.34 |
| Freshwater drum | 2.2 | 2.7 | 1.5 | 7.6 | 2.7 | 2.4 | 3.3 | 3.0 | 1.4 | 2.6 | 4.6 | 3.09 |
| Common carp | 5.5 | 9.9 | 4.6 | 6.1 | 2.7 | 3.2 | 2.6 | 5.0 | 1.9 | 2.3 | 4.8 | 4.42 |
| Goldf ish | 0.8 | 0.3 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.14 |
| Gizzard shad | 0.1 | 0.2 | 1.7 | 0.1 | 0.1 | 0.4 | 0.0 | 0.4 | 0.1 | 0.2 | 0.2 | 0.33 |
| Longnose gar | 0.0 | 0.0 | 0.1 | 0.1 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.07 |
| Bowfin | 0.1 | 0.7 | 0.6 | 0.5 | 0.3 | 0.7 | 0.4 | 0.8 | 0.1 | 0.2 | 0.1 | 0.40 |
| Mooneye | 0.1 | 0.0 | 0.0 | 0.0 | $<0.1$ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.01 |
| Quillback | 0.2 | 0.2 | 0.3 | 0.3 | 0.6 | 1.0 | 0.4 | 1.0 | 0.6 | 0.6 | 0.6 | 0.51 |
| Stonecat | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.00 |
| Total | 103.5 | 53.4 | 89.6 | 93.4 | 127.0 | 154.8 | 109.4 | 122.6 | 116.3 | 155.0 | 159.5 | 116.77 |

Appendix 3. Mean number of fish caught per trap-net lift in Lake Erie. Catch per effort is number caught in 24 hours.

| Species | Year |  |  |  |  |  |  |  |  |  | Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |  |
| Smallmouth bass | 0.1 | 0.0 | 0.0 | 0.1 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.0 | 0.08 |
| Yellow perch | 377.0 | 320.0 | 669.0 | 512.0 | 146.0 | 257.0 | 129.0 | 156.0 | 40.2 | 174.0 | 278.02 |
| Rock bass | 1.2 | 0.8 | 1.9 | 0.9 | 1.5 | 1.3 | 1.0 | 1.5 | 0.7 | 1.5 | 1.23 |
| White bass | 1.5 | 1.5 | 3.7 | 1.4 | 10.5 | 4.9 | 2.5 | 2.8 | 7.6 | 0.4 | 3.68 |
| White perch | 0.0 | 0.1 | 0.3 | 0.5 | 24.6 | 35.0 | 10.9 | 38.9 | 30.3 | 43.5 | 18.40 |
| Pumpkinseed | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.1 | 0.0 | 0.1 | 0.0 | 0.1 | 0.06 |
| Bluegill | 0.0 | 0.0 | <0.1 | 0.0 | 0.0 | $<0.1$ | <0.1 | $<0.1$ | 0.0 | <0.1 | 0.01 |
| Biack crappie | 0.2 | 0.0 | 0.2 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 | 0.2 | 0.2 | 0.11 |
| Channel catfish | 3.5 | 9.7 | 5.4 | 5.8 | 4.9 | 10.6 | 4.6 | 5.5 | 5.4 | 2.7 | 5.79 |
| Brown bullhead | 0.2 | 1.1 | 1.6 | 1.9 | 1.7 | 4.2 | 2.5 | 1.5 | 4.1 | 0.9 | 1.97 |
| Muskellunge | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | <0.1 | 0.0 | <0.01 |
| White sucker | 7.8 | 8.3 | 7.9 | 12.2 | 8.7 | 6.7 | 10.2 | 33.0 | 10.2 | 7.0 | 11.20 |
| Redhorse | 2.4 | 1.2 | 0.6 | 1.0 | 0.8 | 1.5 | 1.7 | 0.3 | 0.0 | 0.0 | 0.96 |
| Freshwater drum | 37.4 | 66.8 | 14.0 | 42.9 | 13.4 | 23.5 | 25.1 | 30.6 | 25.3 | 9.1 | 28.81 |
| Common carp | 5.1 | 26.1 | 4.7 | 8.2 | 6.9 | 14.9 | 3.5 | 2.0 | 1.9 | 0.6 | 7.39 |
| Goldf ish | 4.8 | 2.4 | 0.3 | 0.4 | 0.4 | 2.5 | 0.6 | 0.2 | 0.1 | 0.0 | 1.17 |
| Gizzard shad | 4.4 | 4.7 | 2.3 | 3.9 | 17.8 | 28.4 | 18.1 | 17.4 | 2.7 | 2.3 | 10.19 |
| Longnose gar | 0.1 | 0.0 | 0.0 | <0.1 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.02 |
| Bowf in | 0.0 | 0.0 | 0.0 | <0.1 | 0.0 | 0.0 | 0.0 | 0.0 | $<0.1$ | 0.0 | $<0.01$ |
| Mooneye | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.00 |
| Quillback | 4.0 | 18.6 | 1.8 | 2.0 | 2.4 | 5.6 | 2.0 | 1.9 | 1.7 | 1.8 | 4.18 |
| Stonecat | <0.1 | <0.1 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 | $<0.1$ | <0.1 | <0.1 | 0.03 |
| Total | 449.7 | 461.4 | 713.6 | 593.3 | 240.1 | 396.3 | 212.0 | 291.9 | 90.4 | 70.2 | 416.04 |

Appendix 4. Mean number of walleye caught pet trap-net day in Lake St. Clair, 1972-85

| Age | Year |  |  |  |  |  |  |  |  |  |  |  |  |  | Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |  |
| 2 | 2.76 | 0.20 | 0.51 | 0.02 | 1.26 | 0.72 | 0.22 | 4.84 | 0.08 | 3.21 | 0.97 | 1.48 | 3.44 | 0.04 | 1.41 |
| 3 | 0.57 | 3.92 | 0.44 | 1.06 | 0.22 | 2.72 | 1.87 | 0.14 | 11.32 | 1.28 | 1.36 | 4.16 | 0.81 | 8.33 | 2.73 |
| 4 | 1.11 | 0.46 | 2.64 | 0.52 | 1.07 | 0.15 | 1.67 | 0.80 | 0.44 | 9.46 | 0.46 | 0.96 | 1.58 | 0.41 | 1.55 |
| 5 | 0.35 | 0.46 | 0.32 | 1.43 | 0.60 | 1.10 | 0.40 | 1.02 | 0.83 | 0.34 | 2.40 | 0.49 | 0.52 | 0.77 | 0.79 |
| 6 | 0.54 | 0.19 | 0.42 | 0.31 | 1.53 | 0.49 | 3.49 | 0.59 | 0.53 | 0.25 | 0.08 | 1.11 | 0.27 | 0.28 | 0.72 |
| 7 | 1.46 | 0.21 | 0.30 | 0.31 | 0.17 | 1.70 | 1.60 | 0.44 | 0.44 | 0.23 | 0.06 | 0.06 | 0.52 | 0.20 | 0.55 |
| 8 | 0.51 | 0.61 | 0.25 | 0.18 | 0.19 | 0.11 | 1.60 | 1.16 | 0.33 | 0.18 | 0.15 | 0.06 | 0.02 | 0.15 | 0.39 |
| 9 | 0.16 | 0.12 | 0.86 | 0.08 | 0.09 | 0.06 | 0.22 | 0.18 | 0.19 | 0.13 | 0.08 | 0.04 | 0.00 | 0.02 | 0.16 |
| 10 | 0.00 | 0.05 | 0.08 | 0.41 | 0.02 | 0.00 | 0.22 | 0.01 | 0.03 | 0.07 | 0.09 | 0.05 | 0.02 | 0.01 | 0.08 |
| 11 | 0.10 | 0.05 | 0.04 | 0.04 | 0.02 | 0.06 | 0.12 | 0.06 | 0.00 | 0.03 | 0.02 | 0.02 | 0.00 | 0.00 | 0.04 |
| 12 | 0.06 | 0.01 | 0.12 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 | 0.02 |
| 13 | 0.00 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | 0.05 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 |
| 14 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | <0.01 |
| Sum | 7.61 | 6.32 | 5.99 | 4.36 | 5.15 | 7.12 | 11.48 | 9.26 | 14.21 | 15.19 | 5.69 | 8.44 | 7.19 | 10.23 | 8.45 |
| Net days | 73 | 111 | 143 | 216 | 105 | 129 | 79 | 141 | 110 | 61 | 84 | 184 | 190 | 194 | 130 |

Appendix 5. Mean number of walleye caught per trap-net day in Lake Erie, 1978-87.

|  | Year |  |  |  |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Age | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | Mean |
| 1 | 0.18 | 0.00 | 0.04 | 0.10 | 0.03 | 0.16 | 0.00 | 0.02 | 0.30 | 0.16 | 0.10 |  |
| 2 | 0.38 | 26.03 | 3.46 | 4.65 | 11.99 | 7.14 | 14.29 | 0.83 | 10.02 | 7.12 | 8.59 |  |
| 3 | 6.29 | 0.58 | 3.69 | 2.46 | 1.84 | 4.06 | 3.17 | 38.02 | 2.22 | 5.39 | 6.77 |  |
| 4 | 3.40 | 4.42 | 0.20 | 2.78 | 1.47 | 0.67 | 2.20 | 2.55 | 27.78 | 1.25 | 4.67 |  |
| 5 | 0.25 | 0.81 | 0.45 | 0.16 | 2.10 | 0.47 | 0.41 | 1.60 | 2.29 | 9.91 | 1.84 |  |
| 6 | 0.32 | 0.12 | 0.14 | 0.44 | 0.29 | 0.73 | 0.23 | 0.68 | 0.83 | 0.77 | 0.45 |  |
| 7 | 0.00 | 0.12 | 0.02 | 0.12 | 0.27 | 0.08 | 0.51 | 0.25 | 0.20 | 0.52 | 0.21 |  |
| 8 | 0.00 | 0.12 | 0.02 | 0.04 | 0.12 | 0.14 | 0.05 | 0.31 | 0.41 | 0.19 | 0.14 |  |
| 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.02 | 0.05 | 0.15 | 0.11 | 0.15 | 0.05 |  |
| 10 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.04 | 0.02 | 0.01 | 0.01 |  |
| 11 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 |  |
| 12 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 |  |
| Total | 10.82 | 32.19 | 8.03 | 10.77 | 18.13 | 13.49 | 20.91 | 44.45 | 44.22 | 25.48 | 22.85 |  |
| Net days | 130 | 87 | 97 | 71 | 84 | 106 | 85 | 85 | 70 | 73 | 89 |  |

Appendix 6. Mean number of walleye caught per gill-net gang in Lake Erie, 1978-87

| Age | Year |  |  |  |  |  |  |  |  |  | Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |  |
| 1 | 91.00 | 19.00 | 44.00 | 43.00 | 33.50 | 29.00 | 4.50 | 69.75 | 98.00 | 96.75 | 52.85 |
| 2 | 7.00 | 37.00 | 25.00 | 13.50 | 21.50 | 21.25 | 91.75 | 12.00 | 34.25 | 42.50 | 30.58 |
| 3 | 25.75 | 5.25 | 22.67 | 6.00 | 5.00 | 14.50 | 7.75 | 95.75 | 4.00 | 20.50 | 20.72 |
| 4 | 8.25 | 10.50 | 2.83 | 9.00 | 5.50 | 4.25 | 5.00 | 3.75 | 44.25 | 5.00 | 9.83 |
| 5 | 0.50 | 3.50 | 3.50 | 1.00 | 5.00 | 2.50 | 2.25 | 5.25 | 2.75 | 28.50 | 5.48 |
| 6 | 0.75 | 0.25 | 0.33 | 2.00 | 1.50 | 2.50 | 1.75 | 2.00 | 2.25 | 2.25 | 1.56 |
| 7 | 0.00 | 0.00 | 0.17 | 1.50 | 0.50 | 0.25 | 3.00 | 0.50 | 0.50 | 0.50 | 0.69 |
| 8 | 0.00 | 0.00 | 0.17 | 0.00 | 0.00 | 0.00 | 0.00 | 0.50 | 1.25 | 0.50 | 0.24 |
| 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.50 | 0.50 | 0.25 | 0.00 | 0.13 |
| Total | 133.25 | 75.50 | 98.67 | 76.00 | 72.50 | 74.25 | 116.50 | 190.00 | 187.50 | 196.50 | 122.07 |

Appendix 7. Average annual harvest by walleye anglers in the Lake Erie-Lake St. Clair system, tag recovery ratios, stock composition estimates (from tag recovery ratios), and approximate composition of the harvest by numbers.

|  | Average <br> annual <br> harvest | Tag <br> recovery <br> ratio $^{1}$ | Percent <br> Lake Erie $^{2}$ | Lake Erie <br> walleye | Lake St. Clair <br> walleye |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Lake Erie | 7 million | $740 / 776$ | 95 | 7 million | trace |
| Lake St. Clair | 132,000 | $60 / 782$ | 8 | 10,560 | 121,440 |
| Northern | 120,000 | - | 23 | 27,600 | 92,400 |
| St. Clair River | 164,000 | $167 / 239$ | 70 | 114,800 | 49,200 |
| Detroit River |  |  |  | 7 million | 263,000 |
| Total |  |  |  |  |  |

${ }^{1}$ The ratio of the number of walleye harvested in a particular area that were tagged in Lake Erie to the total number of tags recovered in that area.
${ }^{2}$ Derived from the tag recovery ratio, except that the northern St. Clair River estimate is from the genetic analysis (see text).

Appendix 8. Number of walleye with reward tags and the number reported by anglers in subsequent years from Lake St. Clair.

|  | Reward tag value |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Recovery <br> year | $\$ 2.00$ | $\$ 4.00$ | $\$ 6.00$ | $\$ 8.00$ | Reward | Non-reward |
| 1981 Tags |  |  |  |  |  |  |
| 1981 | 5 | 6 | 9 | 4 | 24 | 22 |
| 1982 | 4 | 2 | 3 | 4 | 13 | 22 |
| 1983 | 1 | 3 | 0 | 1 | 5 | 6 |
| 1984 | 0 | 0 | 0 | 3 | 3 | 3 |
| 1985 | 1 | 2 | 0 | 1 | 4 | 2 |
| 1986 | 0 | 1 | 0 | 0 | 1 | 1 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 |
| Number tagged | 79 | 77 | 80 | 81 | 317 | 650 |
| Total number recovered | 11 | 14 | 12 | 13 | 50 | 56 |
| Percent recovery | 13.92 | 18.18 | 15.00 | 16.05 | 15.77 | 8.62 |


| 1982 Tags |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 3 | 3 | 9 | 5 | 20 | 15 |
| 1983 | 1 | 3 | 2 | 2 | 8 | 10 |
| 1984 | 1 | 2 | 1 | 1 | 5 | 9 |
| 1985 | 1 | 4 | 0 | 0 | 5 | 1 |
| 1986 | 1 | 0 | 0 | 0 | 1 | 3 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 |
| Number tagged | 50 | 48 | 51 | 49 | 198 | 259 |
| Total number recovered | 7 | 12 | 12 | 8 | 39 | 38 |
| Percent recovery | 14.00 | 25.00 | 23.53 | 16.33 | 19.70 | 14.67 |

1983 Tags

| 1983 | 12 | 12 | 9 | 17 | 50 | 94 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1984 | 3 | 3 | 8 | 5 | 19 | 40 |
| 1985 | 1 | 3 | 2 | 3 | 9 | 9 |
| 1986 | 0 | 1 | 1 | 1 | 3 | 3 |
| 1987 | 1 | 0 | 0 | 1 | 2 | 5 |
| Number tagged | 95 | 100 | 100 | 99 | 394 | 1,134 |
| Total number recovered | 17 | 19 | 20 | 27 | 83 | 151 |
| Percent recovery | 17.89 | 19.00 | 20.00 | 27.27 | 21.07 | 13.32 |

Appendix 9. Population statistics for western Lake Erie walleye from the CAGEAN model.
$\left.\begin{array}{cccc}\hline & \begin{array}{c}\text { Instantaneous } \\ \text { fishing } \\ \text { mortality } \\ \text { Fate }\end{array} & \begin{array}{c}\text { Annual } \\ \text { survival } \\ \text { rear }\end{array} & 0.3325\end{array} \begin{array}{c}\text { Commercial } \\ \text { exploitation } \\ \text { rate }\end{array}\right]$

| Fishing <br> year | Angler <br> exploitation <br> rate | Numerical <br> abundance | Numerical <br> catch |
| :---: | :---: | ---: | :---: |
| 1978 | 0.2263 | $6,508,554$ | $1,678,292$ |
| 1979 | 0.1840 | $14,545,434$ | $3,266,564$ |
| 1980 | 0.1702 | $14,115,936$ | $3,359,890$ |
| 1981 | 0.2861 | $12,464,676$ | $4,586,210$ |
| 1982 | 0.2290 | $13,212,614$ | $3,956,338$ |
| 1983 | 0.1022 | $12,839,993$ | $2,492,817$ |
| 1984 | 0.0796 | $47,090,436$ | $6,255,436$ |
| 1985 | 0.1022 | $34,934,950$ | $5,779,240$ |
| 1986 | 0.1035 | $35,556,303$ | $5,493,318$ |
| 1987 | 0.1008 | $41,950,221$ | $5,794,971$ |


[^0]:    ${ }^{1}$ A contribution from Dingell-Johnson Project F-53-R, Michigan.
    ${ }^{2}$ U. S. Fish and Wildlife Service, National Fisheries Research Center-Great Lakes, Ann Arbor, Michigan, USA.

